BRITISH
ECOLOGICAL
SOCIETY

Journal of Ecology doi: 10.1111/1365-2745.12781

Multi-scale drivers of community diversity and composition across tidal heights: an example on temperate seaweed communities

Marine Robuchon*,1,2,3 , Myriam Valero², Eric Thiébaut⁴ and Line Le Gall¹

¹Institut de Systématique, Évolution, Biodiversité, ISYEB UMR 7205 CNRS, MNHN, EPHE, Sorbonne Universités, UPMC Univ Paris 06, Muséum national d'histoire naturelle, 57 rue Cuvier, CP39, 75005 Paris, France; ²Evolutionary Biology and Ecology of Algae, EBEA UMI 3614 CNRS, Sorbonne Universités, UPMC, PUCCh, UACH, Station Biologique de Roscoff, CS 90074, Place Georges Teissier, 29688 Roscoff Cedex, France; ³Unité Biologie des Organismes et Ecosystèmes Aquatiques, BOREA UMR 7208 CNRS, MNHN, 207 IRD, Sorbonne Universités, UPMC Univ Paris 06, UA, UCBN, Muséum national d'histoire naturelle, 43 rue Cuvier, CP26, 75005, Paris, France; and ⁴Sorbonne Universités, UPMC Univ Paris 6, CNRS, Station Biologique de Roscoff, UMR 7144 Adaptation et Diversité en Milieu Marin, CS 90074, Place Georges Teissier, 29688 Roscoff Cedex, France

Summary

- 1. Despite recent advances in understanding community assembly processes, appreciating how these processes vary across multiple spatial scales and environmental gradients remains a crucial issue in ecology.
- 2. This study aimed to disentangle the drivers of diversity and composition of seaweed communities through a gradient of spatial scales based on a hierarchical sampling design consisting of 19 sites distributed in four sectors along the Brittany coastline. Using randomised community matrices and Moran's eigenvector maps (MEMs), we compared (i) the relative importance of deterministic and stochastic processes, (ii) the environmental correlates of community composition, and (iii) the scale of variation in community composition for seaweed communities located at two different tidal heights.
- **3.** Processes shaping community patterns are expected to vary along a gradient of tidal heights. Therefore, we specifically examined the following hypotheses: the contribution of deterministic over stochastic processes as well as the relative importance of environmental filtering over biotic interactions should be enhanced for seaweed communities of the infralittoral fringe compared to subtidal ones, whereas dispersal of propagules in the water column should be more restricted resulting in finer scale variation in community composition for seaweed communities of the infralittoral fringe compared to subtidal communities.
- **4.** Seaweed communities were largely shaped by deterministic processes, although the relative importance of deterministic processes was greater for communities of the infralittoral fringe than for subtidal communities. Sea surface temperature and geophysical variables were correlates of community composition at the two tidal heights; additionally, waves and current were correlated with the composition of the communities of the infralittoral fringe while kelp density was correlated with the composition of subtidal communities. Variation in community composition was observed at a finer scale for infralittoral fringe than for subtidal communities.
- **5.** *Synthesis.* Our results suggest that the relative importance of deterministic and stochastic processes in structuring seaweed communities varies across tidal heights. Furthermore, the Moran's eigenvector maps framework highlights that the nature of environmental correlates and the spatial scale at which they were good correlates of community composition also vary across tidal heights and may therefore be useful to broaden our understanding of community assembly across vertical gradients.

^{*}Correspondence author. Unité Biologie des Organismes et Ecosystèmes Aquatiques, BOREA UMR 7208 CNRS, MNHN, 207 IRD, Sorbonne Universités, UPMC Univ Paris 06, UA, UCBN, Muséum national d'histoire naturelle, 43 rue Cuvier, CP26, 75005 Paris, France. E-mail: robuchon@mnhn.fr

Key-words: Brittany, community composition, determinants of plant community diversity and structure, deterministic and stochastic processes, environmental filtering, kelp, Moran's eigenvector map, multiple spatial scales, seaweed communities, tidal height

Introduction

Understanding how species assemble into communities is a key and highly debated issue in ecology. The structure of communities has been traditionally explained as the result of deterministic processes where species persistence in its environment is determined by abiotic conditions and biotic interactions (Hutchinson 1957; Grime 1973; Tilman 1982). In 2001, Hubbell proposed an alternative view: the unified theory of biodiversity and biogeography which considers the structure of communities as the fruit of stochastic processes only (Hubbell 2001). Over the last 15 years, important efforts have been made to disentangle the relative contribution of deterministic and stochastic processes in shaping community structure, including the proposition of a theoretical framework (Leibold et al. 2004; Logue et al. 2011). It is now widely recognised that both deterministic (i.e. environmental filtering and biotic interactions) and stochastic processes (i.e. ecological drift and limited dispersal) influence community structure (Leibold & McPeek 2006), however, there is no consensus regarding the relative importance of these two kinds of processes which seem to differ both among different environmental conditions (Chase 2007; Chase & Myers 2011) and across spatial scales (Cottenie 2005; Chase & Myers 2011; Logue et al. 2011).

The consideration of multiple spatial scales to better understand patterns of biodiversity and processes driving them has therefore become a cornerstone of modern ecology. Indeed, communities are organised at multiple scales and form a network of communities that are connected by dispersal of multiple potentially interacting species called metacommunities (see Leibold et al. 2004 for review). Connectivity among communities occur at various rates, depending on both species dispersal abilities and landscape features, and affect the structure of metacommunities in interaction with ecological drift, environmental filtering and biotic interactions. Yet, our knowledge regarding how biodiversity is structured across spatial scales varies greatly among ecosystems: most empirical support comes from research on terrestrial ecosystems, in which dispersal distances are much easier to estimate than in marine ecosystems. In the absence of robust dispersal estimates and considering the fluid characteristics of the ocean, it has long been considered that dispersal rates were greater in the marine realm compared to terrestrial environments (Cowen 2000). However, the recent methodological advances in the study of dispersal pathways and connectivity among marine populations have greatly improved estimates of marine dispersal and challenge the simplicity of this long-standing paradigm. As such, in their review of propagule dispersal in marine and terrestrial environments, Kinlan & Gaines (2003) have shown that marine organisms displayed a huge variety

in their mean dispersal distance ranging from a few metres to nearly 1000 km.

In cold to temperate waters, rocky subtidal assemblages are dominated by kelps, brown seaweeds which form underwater forests and are of major ecological importance since they provide habitat, food and protection to a myriad of other marine organisms (Dayton 1985; Steneck et al. 2002). These kelp forests are one of the most diverse and productive ecosystems world-wide (Mann 1973). Biodiversity patterns of kelp-dominated communities have been described mostly at local scales in different places of the world (e.g. Hawkins & Harkin 1985; Leliaert et al. 2000; Christie et al. 2003; Graham 2004; Pehlke & Bartsch 2008; Leclerc et al. 2015), more rarely at regional scales (Wernberg, Kendrick & Phillips 2003; Derrien-Courtel, Le Gal & Grall 2013; Robuchon et al. 2015) but to date, only Smale, Kendrick & Wernberg (2011) looked at these patterns across multiple spatial scales. In their study of the subtidal flora of the south-western Australia coastline, they showed that diversity and turnover of communities varied considerably at all spatial scales, although small-scale variability contributed most to total variation. This important small-scale variability, a common pattern in rocky shore communities (Fraschetti, Terlizzi & Benedetti-Cecchi 2005), was mainly attributed to the action of waves and habitat heterogeneity while regional scale variability was attributed to climatic factors as most species had cool-water affinities (Smale, Kendrick & Wernberg 2011). Despite providing great insights in the understanding of multi-scale variability in kelp-dominated communities, the study of Smale, Kendrick & Wernberg (2011) did not explicitly address the question of what are the underlying processes that explained the biodiversity patterns they documented. Addressing the relative contribution of deterministic and stochastic processes driving biodiversity patterns of kelp-dominated communities across multiple spatial scales remains an open issue.

In recent years, methods to analyse spatial ecological data across different scales have been improved, notably with the emergence of a set of methods now called Moran's eigenvector maps (MEMs; Dray, Legendre & Peres-Neto 2006). These methods can model structures at scales ranging from the broadest down to the finest on the basis of a weighted matrix representing the degree of connection between sampling sites, where the weighted matrix can take several forms from the simplest (a binary matrix: sites are connected or not) to the most geographically realistic (a matrix of geographic distances among sites). A principal coordinates analysis (PCoA) is then performed on the truncated weighted matrix and the resulting eigenvectors that model spatial correlation are used as spatial explanatory variables in canonical ordination (Borcard, Gillet & Legendre 2011). Therefore, the MEMs framework is a way to evaluate the importance of measured explanatory variables in driving community patterns through a gradient of spatial scales as well as to identify significant residual spatial patterns that could arise from the omission of important unmeasured explanatory variables or processes (Dray et al. 2012).

In this study, our objective was to disentangle drivers of community diversity and composition across multiple spatial scales for kelp-dominated seaweed communities along c. 500 km of the Brittany coastline (France). This region harbours a hot spot of seaweed diversity (Kerswell 2006; Keith, Kerswell & Connolly 2014) and forms along the European Atlantic coastline a transition zone between two biogeographic provinces, the warm temperate Lusitanian province in the south and the cold temperate Northern European Seas province in the north (Spalding et al. 2007). We addressed this question in seaweed communities located at two different tidal heights: (i) Laminaria digitata understorey communities, spanning the lower intertidal (i.e. the infralittoral fringe), and upper subtidal zones and (ii) Laminaria hyperborea communities, found in the subtidal zone (i.e. the infralittoral zone). Note that, even if L. digitata understorey communities are located both in the infralittoral fringe and the upper subtidal zone, they will be referred hereafter as infralittoral fringe communities to facilitate the reading. Based upon these contrasted tidal heights, we can formulate three hypotheses regarding the differences expected between infralittoral fringe communities and subtidal ones in terms of relative importance of deterministic processes, environmental drivers of community composition and propagule dispersal distances (Fig. 1). First, some recent works have suggested that the relative importance of deterministic processes in structuring communities was greater in disturbed compared to undisturbed environment (Chase 2007; Chase & Myers 2011). The intertidal environment experiences frequent changes between immersion and emersion and is exposed to waves; hence it is more disturbed than the subtidal one which is always immersed (Raffaelli & Hawkins 1996). We can, therefore, hypothesise that the relative importance of deterministic processes in structuring communities is greater in infralittoral fringe communities than in subtidal ones. Second, the relative importance of environmental filters over biotic interactions in structuring littoral communities is known to increase with tidal height (Raffaelli & Hawkins 1996). Therefore, the relative importance of abiotic over biotic variables in driving community composition is expected to be greater in infralittoral fringe communities compared to subtidal ones. Finally, third, infralittoral fringe communities are located higher on the vertical gradient of tidal heights than subtidal communities. Consequently, dispersal distances should be lower in these communities because they are less often immersed (which limits the dispersal of propagules in the water column) but most importantly they experience osmotic and thermic stresses during periods of emersion (which stimulate the simultaneous release of propagules at low tides and thus short-distance dispersal, Norton 1992). Such differences in terms of propagule dispersal between the infralittoral fringe and the subtidal zone have recently been

evidenced by comparing genetic connectivity of L. digitata and L. hyperborea populations (Robuchon et al. 2014). Therefore, variation in community composition should be observed at a finer spatial scale in infralittoral fringe communities compared to subtidal ones.

Towards our objective, we conducted an extensive and quantitative survey of seaweed communities using a nested sampling design, characterised the variation in diversity indices and their deviations from null models at the different levels of our sampling hierarchy and investigated how environmental variables fitted community composition at multiple spatial scales using the MEMs framework.

Materials and methods

STUDY AREA AND DATA SOURCES

To determine the drivers of seaweeds' community structure across spatial scales, we compiled data on floristic composition and environmental variables across 19 sites distributed in four sectors of Brittany (France, Fig. 2). These sectors were chosen because they display distinct features which characterised the environmental heterogeneity of the Brittany coastline: St Malo Bay shows a more irregular topography and is characterised by the presence of cyclonic and anticyclonic gyres that increase the water mass residence times and may affect propagule dispersal (Salomon & Breton 1993), water bodies of Southern Brittany are stratified (Le Fèvre 1987) and between these two regions, Iroise Sea and Morlaix Bay form a cold and resilient water pocket (Gallon et al. 2014). Floristic composition was assessed during a survey of seaweed diversity conducted in winter 2011 by scuba diving and targeting the flora living beneath the canopy of L. digitata and L. hyperborea, which differ by their distributions along the tidal zone: L. digitata occupies the infralittoral fringe, between +1 and -1 m depth whereas L. hyperborea occupies the infralittoral zone between -1 and -30 m (reviewed by Robuchon et al. 2014). This difference in vertical distribution implies that L. digitata populations and associated understorey communities are sometimes emerged (83 h over the year 2011 based on the SHOM data - http://www.shom.fr/ - for the city of Roscoff, in Morlaix Bay), whereas L. hyperborea populations and associated understorey communities are always underwater. At each site, six quadrats of 0·10 m² were randomly placed among the kelps holdfasts at a few metres of distance (three among L. digitata and three among L. hyperborea) and sampled for all macroscopic specimens of seaweeds (except crustose seaweeds) present in these quadrats. Then, specimens were sorted by morphotype and identified using the floristic keys and field guides available for the region (Dixon & Irvine 1977; Irvine 1983; Fletcher 1987; Burrows 1991; Maggs & Hommersand 1993; Irvine, Chamberlain & Maggs 1994; Brodie & Irvine 2003; Cabioc'h et al. 2006). The number of individuals per morphotype within each quadrat was counted, allowing generating one quadrat-by-species and one site-by-species abundance matrices.

We also built a site-by-environment matrix resulting from the compilation of 32 environmental variables (Table 1) related to the density of Laminaria individuals (measured during the floristic survey), geophysical and bioclimatic characteristics (extracted from MARSPEC layers, Sbrocco & Barber 2013) and sea-states characteristics (calculated from the HOMERE database, Boudière et al. 2013).

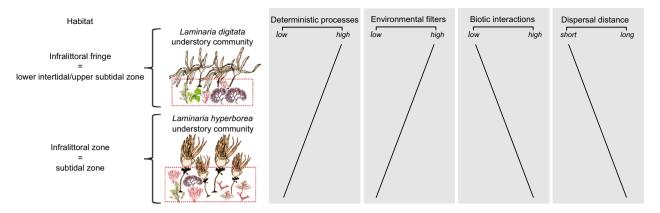


Fig. 1. Diagram showing the different hypotheses of our study regarding the relative importance of deterministic processes, environmental filters, biotic interactions and dispersal distance in structuring seaweed communities at two tidal heights.

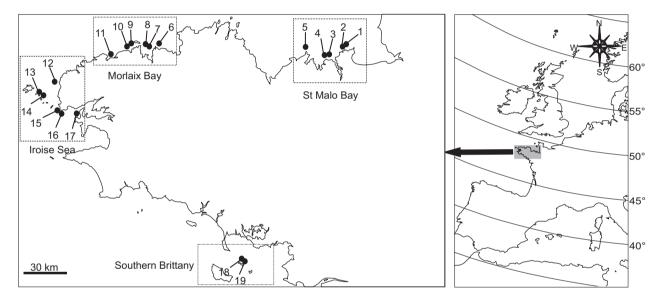


Fig. 2. Map showing the 19 sampling sites distributed in the four sectors along the Brittany coastline, France. Sites are: Guimereux (1), La Bigne (2), Nerput (3), Le Moulin (4), Les Amas du Cap (5), Primel (6), Duons Est (7), Duons Ouest (8), Santec 2 (9), Santec 1 (10), Les Amiettes (11), Les Linious (12), Men Vriant (13), Klosenn Malaga (14), Les Rospects (15), St Mathieu (16), Pointe du Grand Gouin (17), Houat 2 (18) and Houat 1 (19). For site coordinates, see Appendix S1, Supporting Information.

HIERARCHICAL ANALYSES OF DIVERSITY INDICES

Species richness (SR) and Shannon diversity index based on log_e (H', Shannon 1948) were calculated for the infralittoral fringe communities and the subtidal communities at three spatial scales: sector (4 levels), site (19 levels) and quadrat (57 levels). Variability in SR and H' was examined at the different spatial scales with nested ANOVA conducted with the R package BiodiversityR (Kindt & Coe 2005). The design was fully hierarchical; sectors were fixed while nested factors were treated as random. As neither SR nor H' fit the assumptions of normality and homogeneity of variance, a dissimilarity matrix based on Bray–Curtis coefficients (Bray & Curtis 1957) derived from untransformed SR and H' (using a dummy variable equal to 1) was generated for the analyses, which used 999 permutations.

To examine if observed SR and H' at the three spatial scales of sampling hierarchy differed from results expected under a null model, we performed an additive diversity partitioning following Crist $et\ al.$ (2003) where mean values of α diversity at lower levels of a sampling hierarchy are compared to the total diversity in the entire dataset, that

is, γ diversity. The expected diversity components were calculated 999 times by individual-based randomisation of the community data matrix using the R package vegan (Oksanen *et al.* 2015).

MULTIVARIATE SPATIAL ANALYSES

To analyse spatial structures of seaweed communities across multiple spatial scales, we used the approach proposed by Dray et al. (2012). It consists of examining how the spatial pattern of beta diversity changes when considering the initial site-by-species abundance matrix (i.e. the community matrix), its approximation by environmental variables (i.e. the fitted matrix) and its residual counterpart (i.e. the residual matrix) on the one hand (McIntire & Fajardo 2009) and to estimate and test at which spatial scale these beta diversity changes occur using MEMs (Dray, Legendre & Peres-Neto 2006) on the other hand.

We carried out this approach independently for infralittoral fringe communities and for subtidal communities. To that purpose, we

(continued)

Table 1. Characteristics of the environmental variables used in this study

Name (abbreviation)	Description (units)	Type (resolution)	Source
Density of Laminaria digitata	Density of L. digitata individuals [semi-quantitative measure ranging	In situ measure	This study
(La.density)* Density of <i>Laminaria hyperborea</i> (Lh density) [†]	from 1 (low density) to 4 (nign density)] Density of <i>L. hyperborea</i> individuals [semi-quantitative measure ranging from 1 (low density) to 4 (high density)]	In situ measure	This study
Bathymetry (bathy) East/West aspect (ew.aspect)**	Depth of the seafloor (m) Horizontal orientation of the seafloor on the East/West gradient (radians)	Remotely sensed measure (30 arc-second) Derived from remotely sensed bathymetry	MARSPEC [‡] MARSPEC [‡]
North/South aspect (ns.aspect)*·†	Horizontal orientation of the seafloor on the North/South gradient (radians)	(50 arc-second) Derived from remotely sensed bathymetry (30 arc-second)	$MARSPEC^{\ddagger}$
Plan curvature (pl.curv)*.↑	Terrain curvature in the direction perpendicular to the maximum slope (none)	Derivation remotely sensed bathymetry (20 arc-second)	$MARSPEC^{\ddagger}$
Profile curvature (pr.curv)	Terrain curvature in the direction parallel to the maximum slope (none)	(20 arc-second) Derived from remotely sensed bathymetry (30 arc-second)	MARSPEC*
Distance to shore (shore.dist)**†	Distance to shore (km)	Deriver are second. (30 arc-second)	$MARSPEC^{\ddagger}$
Bathymetric slope (slope)**†	Slope of the seafloor (°)	Deriver and sensed bathymetry (30 arc-second)	MARSPEC [‡]
Concavity (concav)	Slope of the bathymetric slope (°)	Deriver second (20 are second)	MARSPEC [‡]
Mean annual SSS (sss.mean)	Mean sea surface salinity averaged over the period 1955–2006 (psu)	Derived from in situ measures (30 arc-second)	$MARSPEC^{\ddagger}$
Minimum monthly SSS (sss.min)	Salinity of the least salty month averaged over the period 1955–2006 (psu)	Derived from <i>in situ</i> measures (30 arc-second)	MARSPEC*
Maximum monthly 555 (sss.max) Annual range in SSS (sss range)	Saminy of the saftlest month averaged over the period 1933–2000 (psu) Annual range in sea surface salinity averaged over the neriod 1955–2006 (nsu)	Derived from <i>in situ</i> measures (30 arc-second)	MARSPEC*
Annual variance in SSS (sss.var)	Annual variance in sea surface salinity averaged over the period 1955–2006 (psu)	Derived from in situ measures (30 arc-second)	MARSPEC [‡]
Mean annual SST (sst.mean)**	Mean sea surface temperature averaged over the period 1955-2006 (°C)	Derived from remotely sensed measures (30 arc-second)	MARSPEC*
Minimum monthly SST (sst.min) Maximum monthly SST (sst max)	Temperature of the coldest month averaged over the period 1955–2006 (°C) Temperature of the warmest month averaged over the period 1955–2006 (°C)	Derived from remotely sensed measures (30 arc-second)	MARSPEC*
Annual range in SST (sst.range)**	Annual range in sea surface temperature averaged over the period 1955–2006 (°C)	Derived from remotely sensed measures (30 arc-second)	MARSPEC*
Annual variance in SST (sst.var)	Annual variance in sea surface temperature averaged over the period 1955–2006 (°C)	Derived from remotely sensed measures (30 arc-second)	$MARSPEC^{\ddagger}$
Mean annual CGE (cge.mean)*	Mean wave energy flux averaged over the period 1994–2012 (kW m^{-1})	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study
Minimum monthly CGE (cge.min)	Wave energy flux of the energy-calmest month averaged over the period $1994-2012$ (kW m ⁻¹)	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study
Maximum monthly CGE (cge.max) [†]	Wave energy flux of the energy-most agitated month averaged over the period 1994–2012 (kW m ⁻¹)	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study
Annual range in CGE (cge.range)	Annual range in wave energy flux averaged over the period 1994–2012 (kW m ⁻¹)	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study
Mean annual number of CGE extreme measures (cge.next.mean)	Mean number of measures in wave energy flux exceeding 14 kW $\rm m^{-1}$ averaged over the period 1994–2012 (none)	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study

Table 1. (continued)

Name (abbreviation)	Description (units)	Type (resolution)	Source
Number of CGE extreme measures	Number of measures in wave energy flux exceeding 14 kW m ⁻¹ within	Derived from model outputs in HOMERE database [§]	This study
within the 3 months preceding the	the 3 months preceding the floristic survey (none)	(value from the nearest node)	
floristic survey (cge.next.3 months)	11-2> 0100 MOT L	Desired from the latest and the late	
Mean annual COR (cur.mean)	Mean sea water velocity averaged over the period 1994–2012 (III s.)	(value from the nearest node)	I IIIS SWAY
Minimum monthly CUR (cur.min)*:†	Sea water velocity of the current-calmest month averaged over the period 1994–2012 (m $\rm s^{-1})$	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study
Maximum monthly CUR (cur.max)	Sea water velocity of the current-most agitated month averaged over the period 1994–2012 (m $\rm s^{-1})$	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study
Annual range in CUR (cur.range)	Annual range in sea water velocity averaged over the period 1994–2012 (m $\rm s^{-1})$	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study
Mean annual number of extreme CUR measures (cur.next.mean)	Mean number of measures in sea water velocity exceeding $0.15~\mathrm{m~s^{-1}}$ averaged over the period 1994–2012 (none)	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study
Number of CUR extreme measures within the 3 months preceding the floristic survey (cur.next.3 months)	Number of measures in sea water velocity exceeding $0.15~\mathrm{m~s^{-1}}$ within the 3 months preceding the floristic survey (none)	Derived from model outputs in HOMERE database [§] (value from the nearest node)	This study

*Variable retained in E_Ld.

†Variable retained in E_Lh.

*Sbrocco & Barber (2013).

*Boudière *et al.* (2013).

considered two initial site-by-species abundance matrices, one for infralittoral fringe communities and one for subtidal communities as well as the two corresponding site-by-environment matrices containing explanatory environmental variables (i.e. the environmental matrices). The environmental matrices were generated by testing correlations among all variables listed in Table 1 and removing variables driving absolute values of pairwise correlations superior to 0.75 using the R package caret (Kuhn 2015). The community matrices were transformed using the Hellinger transformation to put emphasis on abundant species as recommended by Rao (1995). Then, we performed a principal component analysis (PCA) to identify the main patterns in community data.

Redundancy analysis (RDA) was conducted to reveal the main structures in community data explained by environmental variables (i.e. analysis of the two fitted matrices) using a forward selection procedure to retain informative environmental variables only, whereas partial residual analysis (PRA) was performed to identify the structures in community data not explained by environmental variables (i.e. analysis of the two residual matrices). To generate the spatial explanatory variables, we conducted a classical distance-based MEMs approach (formally called Principal Coordinates of Neighbour Matrices) following guidelines in Borcard, Gillet & Legendre (2011): (i) we computed a matrix of geographic distances among sites using the shortest path by the sea, (ii) we truncated the matrix using a distance threshold equal to the maximum distance between two consecutive sites across the coastline, (iii) we performed a PCoA on the truncated distance matrix and (iv) we retained 18 eigenvectors that model spatial correlation as spatial explanatory variables. Scalograms were computed for the community, the fitted and the residual matrices by projecting the sites scores on the first two axes of the different analyses (PCA, RDA and PRA respectively) onto the spatial basis formed by the 18 MEMs, therefore representing a partitioning of the respective variances across multiple spatial scales ranked from the broadest to the finest. They are represented in a smoothed version (as in Munoz 2009) with six spatial components formed by groups of three successive MEMs, which is a way to avoid undesired sampling artefacts at fine scales (aliasing effects, Platt & Denman 1975). The individual R² values that form the scalograms and correspond to the amount of variation explained by a given scale are expected to be uniformly distributed in the absence of spatial structure (Ollier, Couteron & Chessel 2006). To uncover significant spatial structure, we therefore tested if the maximum observed R² was significantly larger than values obtained in the absence of spatial structure using a permutation procedure with 999 repetitions.

All analyses were carried out in R (R Core Team 2015) and based on the script provided in Dray et al. (2012) and adapted to our data.

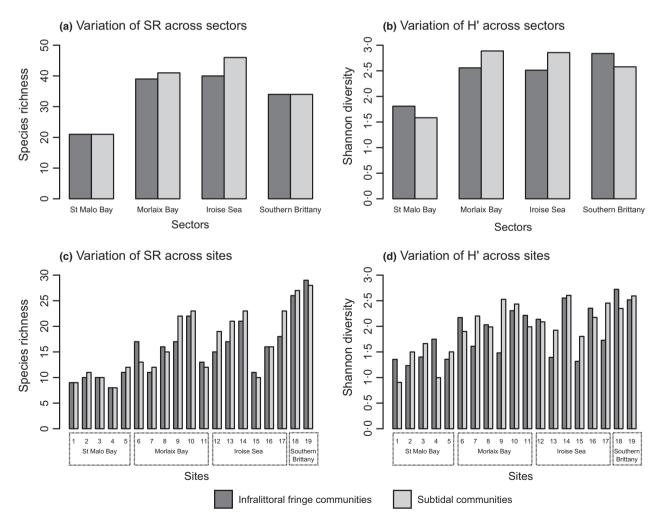


Fig. 3. Comparison of species diversity estimates between infralittoral fringe communities and subtidal communities across space. (a) Species richness (SR) by sector, (b) Shannon diversity (H') by sector, (c) SR by site and (d) H' by site. For site names, see Fig. 2.

Results

In total, 5292 specimens corresponding to 69 species were identified from the 57 quadrat samples of the infralittoral fringe communities and 5179 specimens corresponding to 68 species were identified from the 57 quadrat samples of the subtidal communities. For both infralittoral fringe and subtidal communities, red seaweeds constituted the most abundant and the most diverse seaweed lineage (respectively 94% and 92% of total abundance/78% and 76% of total SR), followed by brown seaweeds (respectively 3.9% and 7.2% of total abundance/14% and 18% of total SR) and green seaweeds (respectively 0.2% and 0.1% of total abundance/5.8% and 4.4% of total SR).

At the scale of sector and for both infralittoral fringe and subtidal communities, SR was minimum in St Malo Bay and maximum in Iroise Sea (equally maximum in Morlaix Bay for infralittoral fringe communities, Fig. 3a). Shannon diversity (H') was also minimum in St Malo Bay for both infralittoral fringe and subtidal communities; however, it reached its maximum in Southern Brittany for infralittoral fringe communities and in Morlaix Bay for subtidal communities (Fig. 3b). At the scale of site and for both communities, SR encountered its minimum in site 4 (a site in St Malo Bay) and its maximum in site 19 (a site in Southern Brittany), and could be markedly variable among sites within the same sector as reported in Morlaix Bay and Iroise Sea (Fig. 3c). Finally, H' was minimum in sites of St Malo Bay and very variable both among sites and between kelp communities (Fig. 3d). For both infralittoral fringe and subtidal communities, nested ANOVA indicated that both SR and H' were considerably and significantly variable at all spatial scales (Table 2). In all

Table 2. Results of nested ANOVA for factor 'site' nested in factor 'sector' based on Bray-Curtis dissimilarity coefficients derived from untransformed species richness (SR) and Shannon diversity (H') using a dummy variable equal to 1; all tests used 999 permutations

					Var. comp
	d.f.	SS	F	P	(%)
(a) SR, Infralitt	oral fring	ge commun	ities		
Sector	3	0.942	9.64	0.001	44.9
Sector:Site	15	0.489	1.85	0.040	23.3
Residuals	38	0.669	0.02		31.9
(b) H', Infralitto	oral fring	ge communi	ities		
Sector	3	0.198	5.82	0.004	36.5
Sector:Site	15	0.170	2.45	0.009	31.3
Residuals	38	0.175	< 0.01		32.2
(c) SR, Subtida	l commu	ınities			
Sector	3	0.778	6.37	0.006	40.9
Sector:Site	15	0.611	3.02	0.005	32.1
Residuals	38	0.512	0.01		26.9
(d) H', Subtidal	commu	nities			
Sector	3	0.196	6.16	0.006	36.4
Sector:Site	15	0.159	2.20	0.030	29.6
Residuals	38	0.183	< 0.01		34.0

d.f., degrees of freedom; SS, sum of squares; F, F-ratio; P, P-value estimating the significance of F-ratios; Var.comp, estimated relative contribution to total variance.

cases, the examination of variance components suggested that the relative importance of fine-scale variability (i.e. site and residuals) outweighed that of broad-scale variability (i.e. sector), although the relative importance of broad-scale variability was higher for SR than for H'.

Results of additive diversity partitioning indicated that in almost all cases, observed SR and H' were significantly different from expectations under a null model for α and β diversities and at the three levels of sampling hierarchy (quadrat, site and sector) (Table 3). Specifically, α diversity was always lower than expected. The only exception was found for β diversity in subtidal communities for which among-sites differences in terms of SR did not differ from null expectations.

Regarding the computation of the environmental matrices, 10 out of 32 environmental variables were retained for the

Table 3. Results of additive diversity partitioning from untransformed species richness (SR) and Shannon diversity (H') comparing simulated statistics under a null model to observed statistics of α and β diversity at three levels of sampling hierarchy ('quadrat', 'site' and 'sector') as well as γ diversity of the entire dataset; all tests used 999 simulations

	Observed statistic	SES	Mean simulated statistic	P
(a) SR Infralit	toral fringe com	munities		
α. quadrat	9.05	-72·30	25.02	0.001
α. site	15.63	-56·61	37.53	0.001
α. sector	33.5	-31.58	52.55	0.001
γ	61.00	0.00	61.00	1.000
β. quadrat	6.58	-16.16	12.49	0.001
β. site	17.87	4.36	15.07	0.015
β. sector	27.50	31.58	8.50	0.001
(b) H', Infralitt	oral fringe com	munities		
α. quadrat	1.55	-91.00	2.66	0.001
α. site	1.88	-89.98	2.90	0.001
α. sector	2.43	-155.69	3.02	0.001
γ	3.03	0.00	3.03	1.000
β. quadrat	0.33	7.89	0.24	0.001
β. site	0.55	38.71	0.11	0.001
β. sector	0.61	155.69	0.02	0.001
(c) SR, Subtida	al communities			
α. quadrat	9.81	-63.83	22.36	0.001
α. site	16.53	-47.11	35.10	0.001
α. sector	35.50	-27.12	53.10	0.001
γ	62.00	0.00	62.00	1.000
β. quadrat	6.72	-16.67	13.42	0.001
β. site	18.97	1.34	19.32	0.171
β. sector	26.50	27.12	10.25	0.001
(d) H', Subtida	l communities			
α. quadrat	1.64	-76.16	2.57	0.001
α. site	1.98	-74.74	2.79	0.001
α. sector	2.48	-130.93	2.90	0.001
γ	2.92	0.00	2.92	1.000
β. quadrat	0.34	12.74	0.22	0.001
β. site	0.50	33.42	0.11	0.001
β. sector	0.44	130-93	0.02	0.001

SES, standardised effect sizes of the observed statistic quantifying the size of the difference between expected and observed values; *P*, *P*-value of the statistic based on simulations.

infralittoral fringe as well as for the subtidal zone of which eight are common to the two matrices (Table 1). After a forward selection procedure to retain environmental variables which best explained the variations of the community matrices, 4 out of the 10 previously retained environmental variables were selected for explaining the infralittoral fringe community matrix (i.e. bathymetry, maximum monthly sea surface temperature, maximum monthly wave energy flux and mean annual sea water velocity) and four out of the eight previously retained environmental variables were selected for explaining the subtidal community matrix (i.e. density of L. hyperborea, distance to shore, bathymetric slope, annual range in sea surface temperature).

The environmental variables explained a significant proportion of the variation of both the infralittoral fringe community matrix $(R^2 = 0.457, P = 0.001)$ and the subtidal community matrix $(R^2 = 0.461, P = 0.001)$. The fitted matrix of the infralittoral fringe exhibited two prominent axes representing a total of 83.9% of the total variance, correlating mainly with maximum monthly sea surface temperature (r = -0.79 for the first axis and -0.58 on the second axis) and bathymetry (r = -0.29 for the first axis and -0.52 on the second axis).Representing a total of 78.0% of the total variance, the first two axes of the subtidal fitted matrix correlated mainly with density of L. hyperborea (r = -0.60) for the first axis and -0.78 on the second axis) and annual range in sea surface temperature (r = -0.83 for the first axis and 0.29 on the second axis).

Figures 4-6 show ordination of sites and the associated scalograms of the main ordination axes for the community matrices (Fig. 4), the fitted matrices (Fig. 5) and the residual matrices (Fig. 6). The scalograms for the first two axes exhibited distinct shapes for the infralittoral fringe community matrix and the subtidal community matrix, with variance accumulation in both broad- (axes 1 and 2) and fine-scale (axis 1 only) components for the infralittoral fringe community matrix (Fig. 4a) and accumulation in broad-scale components only for the subtidal community matrix (Fig. 4b). Indeed, the first axis of the infralittoral fringe community matrix exhibited a fine-scale non-random spatial pattern $(R^2_{\text{Max}} = 0.53, P = 0.008)$ and an important but nonsignificant broad-scale component ($R^2_{\text{Max}} = 0.38$, P = 0.073) while its second axis showed a broad-scale non-random spatial pattern $(R^2_{\text{Max}} = 0.47, P = 0.007)$. In contrast, the main axis of the subtidal community matrix exhibited significantly skewed distributions towards the broad-scale components solely (axis 1: $R^2_{\text{Max}} = 0.79$, P = 0.001; axis 2: $R^2_{\text{Max}} = 0.44$, P = 0.023). These results regarding community matrices signify that community variability is important at both broad and fine scales for infralittoral fringe communities and at broadscale only for subtidal communities.

Regarding the fitted matrices (Fig. 5), scalograms of the first two axes displayed a broad-scale non-random spatial pattern for both the subtidal fitted matrix (axis 1: $R^2_{\text{Max}} = 0.47$, P = 0.011; axis 2: $R^2_{\text{Max}} = 0.68$, P = 0.004) and the infralittoral fringe fitted matrix (axis 1: $R^2_{\text{Max}} = 0.76$, P = 0.001; axis 2: $R^2_{\text{Max}} = 0.37$, P = 0.048). In addition, the first axis of the infralittoral fringe fitted matrix showed a fine-scale nonrandom spatial pattern ($R^2_{\text{Max}} = 0.47$, P = 0.013) while the second axis of the subtidal fitted matrix exhibited an impornonsignificant medium-scale component $(R^2_{\text{Max}} = 0.32, P = 0.109)$. These results regarding fitted matrices are similar to those observed for the community matrices and indicate that environmental variables well explain community variability at broad spatial scales for both infralittoral fringe and subtidal communities, and, additionally, at fine-scale for infralittoral fringe communities.

Finally, the scalograms for the first two axes of the residual matrices exhibited distinct patterns for the infralittoral fringefitted matrix (Fig. 6a), showing variation accumulated mainly and significantly in broad-scale components (axis 1: $R^2_{\text{Max}} = 0.37$, P = 0.046; axis 2: $R^2_{\text{Max}} = 0.52$, P = 0.005), and the subtidal-fitted matrix (Fig. 6b), displaying variation accumulated in medium-scale components (and only significantly for axis 1: $R^2_{\text{Max}} = 0.41$, P = 0.043). Regarding infralittoral fringe communities, this indicates that a significant broad-scale spatial pattern remained in the data after the effects attributable to the measured environmental variables (mainly a combination of maximum temperature and bathymetry) were partialled out. Regarding subtidal communities, this reveals that a significant medium-scale spatial pattern remained in the data after the broad-scale effects related to the measured environmental variables (mainly temperature range and kelp density) were removed.

Discussion

In marine ecology, the variation in community patterns along the vertical gradient of tidal heights is a long-standing issue (Raffaelli & Hawkins 1996) while its examination along a gradient of multiple spatial scales is more recent and still scarce in major coastal ecosystems such as kelp forests (but see Smale, Kendrick & Wernberg 2011). In this study of kelp-dominated seaweed communities, we sought to understand how the relative roles of stochastic (i.e. ecological drift and limited dispersal) and deterministic (i.e. environmental filtering and biotic interactions) processes in structuring communities vary both along the vertical gradient of tidal heights and across multiple spatial scales. Specifically, we highlight the crucial role of deterministic processes in shaping these communities, we identify environmental correlates of community composition and we show that the scale of variation in community composition differs across tidal heights.

RELATIVE IMPORTANCE OF DETERMINISTIC PROCESSES IN STRUCTURING COMMUNITIES

Our results show that variability in α and β species diversity differed from expectations under a null model at all spatial scales for infralittoral fringe communities, indicating that they are largely shaped by deterministic processes at all scales from the finest (the quadrats, separated by a few metres) to the broadest (the sectors, separated by more than 60 km). The same pattern was observed for subtidal

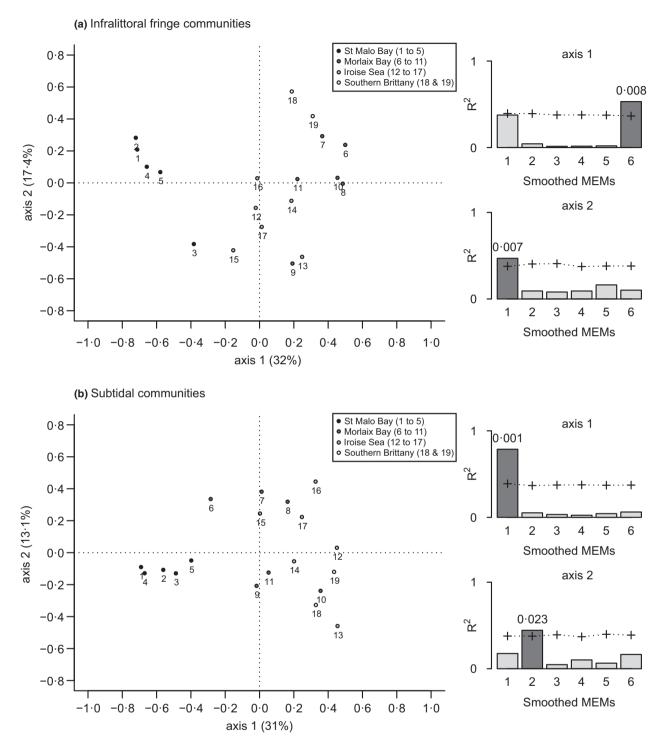


Fig. 4. Plots of site scores on the first two axes of the principal components analyses of site-by-species abundance matrices (a) for infralittoral fringe communities and (b) for subtidal communities. For each score, a smoothed scalogram [the 18 Moran's eigenvector maps (MEMs) are assembled in 6 groups] indicates the portion of variance (R^2) explained by each spatial scale ranked from the broadest to the finest. For each scalogram, the scale corresponding to the highest R^2 (in dark grey) is tested using 999 permutations of the observed values (P-values are given). The 95% confidence limit is also represented by the line of plus signs.

communities, except for the variability in β SR which did not differ from null expectations at the intermediate scale of sites (separated by more than 300 m), suggesting that variability in SR between sites can be explained by the action of stochastic processes only. Therefore, the relative importance of deterministic processes in structuring

communities appears to be slightly greater in infralittoral fringe communities than in subtidal ones. Such differences might be related to the fact that infralittoral fringe communities inhabit a more disturbed environment than subtidal communities: the prevalence of deterministic processes in disturbed environments has been advocated to explain

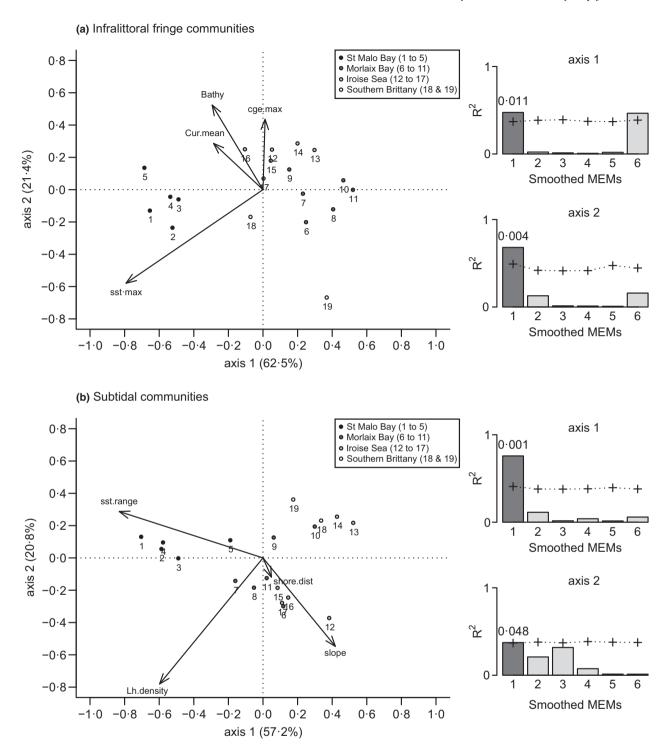


Fig. 5. Plots of site scores on the first two axes of the redundancy analyses using the site-by-environment matrices as predictors (a) for infralittoral fringe communities and (b) for subtidal communities. For each score, a smoothed scalogram [the 18 Moran's eigenvector maps (MEMs) are assembled in 6 groups] indicates the portion of variance (R^2) explained by each spatial scale ranked from the broadest to the finest. For each scalogram, the scale corresponding to the highest R^2 (in dark grey) is tested using 999 permutations of the observed values (P-values are given). The 95% confidence limit is also represented by the line of plus signs.

variation in community patterns along a gradient of disturbance in a variety of other organisms (i.e. small freshwater ponds: Chase 2007 and plants: Myers & Harms 2011). Nevertheless, it does not imply that deterministic processes do not structure subtidal communities. In particular, α diversity was lower than expected for the three spatial

scales and the two diversity metrics examined. These results suggest that distribution of kelp understorey seaweeds is far from random and may be the result of species-specific factors such as biogeographic history and dispersal ability as well as deterministic processes of environmental filtering and/or biotic interactions.

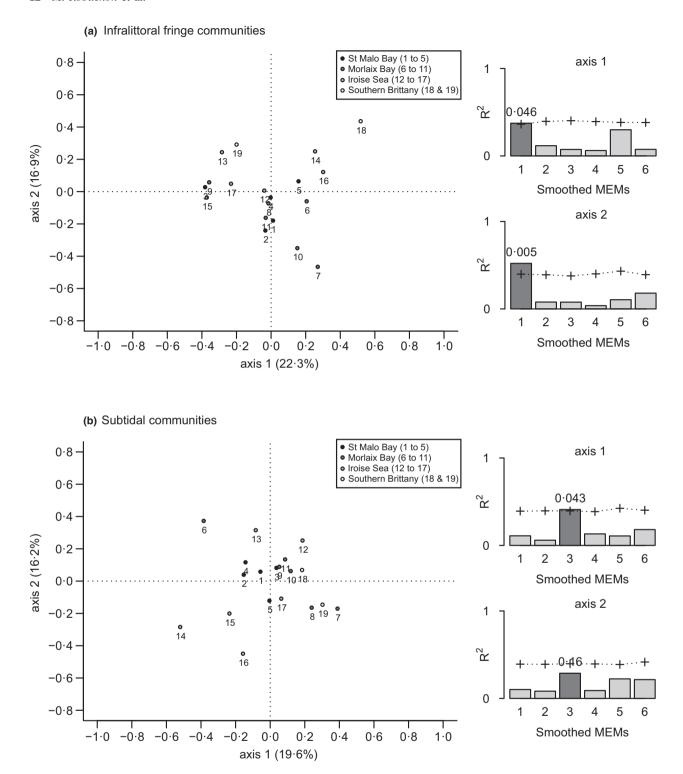


Fig. 6. Plots of site scores on the first two axes of the partial principal component analysis using the site-by-environment matrices as co-variables (a) for infralittoral fringe communities and (b) for subtidal communities. For each score, a smoothed scalogram [the 18 Moran's eigenvector maps (MEMs) are assembled in 6 groups] indicates the portion of variance (R^2) explained by each spatial scale ranked from the broadest to the finest. For each scalogram, the scale corresponding to the highest R^2 (in dark grey) is tested using 999 permutations of the observed values (P-values are given). The 95% confidence limit is also represented by the line of plus signs.

ENVIRONMENTAL CORRELATES OF COMMUNITY COMPOSITION

The relative importance of abiotic over biotic variables in driving community composition established from the list of variables retained in our analysis was slightly greater in infralittoral fringe communities compared to subtidal ones: four abiotic variables correlated with composition of infralittoral fringe communities, whereas a combination of one biotic and three abiotic variables correlated with composition of subtidal communities. Community composition was correlated with variables related to sea surface temperature for both infralittoral fringe and subtidal communities. This finding is consistent with previous studies on kelp forests showing that distribution of understorey red seaweeds was mainly driven by annual amplitude in sea surface temperature (Gallon et al. 2014) and that large-scale community variation was related to the differences in species temperature affinities (Smale, Kendrick & Wernberg 2011; Derrien-Courtel, Le Gal & Grall 2013). In agreement with recent reports documenting the sensitivity of kelp species (e.g. Pehlke & Bartsch 2008; Oppliger et al. 2014) and associated communities (e.g. Gallon et al. 2014; Wernberg et al. 2016) to rising temperatures, our study therefore suggests that kelp forests would be largely affected by climate change.

Geophysical variables were also correlated with community composition and more specifically bathymetry for infralittoral communities and slope for subtidal ones. These results are not surprising since bathymetry and slope might capture the action of other factors (not included in our study) important for the settlement and growth of sessile organisms. For example, light availability decreases with bathymetry and sediment burying decreases with slope and these two factors have been previously described as influencing the composition of rocky subtidal communities (e.g. Miller & Etter 2008).

Moreover, some variables were correlates of composition for infralittoral fringe communities but not for subtidal ones, and vice versa. As such, current and wave energy explained variation in composition for infralittoral fringe communities but not for subtidal ones, a finding which is logical since infralittoral fringe communities are located closer to the sea surface and are therefore more prone to be affected by the action of currents and waves than subtidal communities. A strong hydrodynamism can affect kelp-dominated communities either directly through dislodgement of individual kelps (Wernberg & Connell 2008) or indirectly by modulating herbivores' abundances (Vanderklift, Lavery & Waddington 2009). Furthermore, kelp density explained variation in composition for subtidal communities but not for infralittoral fringe communities. This outcome might reflect an interaction between L. hyperborea canopy and its understorey community. Such canopy-understorey interaction can be either competitive or facilitative. Specifically, canopy formers may competitively exclude understorey species by shading their environment and scouring recruits or juveniles or facilitate the recruitment and existence of other species by mitigating physical stress such as hydrodynamic forces (e.g. Kain 1979; Wernberg, Kendrick & Toohey 2005; Bennett & Wernberg 2014). As we recorded lower species diversity in sites characterised by a high kelp density, the canopy-understorey interactions we detected in subtidal communities are likely dominated by competition.

Our results revealed that the measured environmental variables were good correlates of community composition at both broad and fine scale for infralittoral fringe communities, and at broad scale only for subtidal communities; however, significant spatial patterns remained in the data after the effects attributable to the measured environmental variables were partialled out. These remaining significant spatial patterns could be attributed either to unmeasured environmental variables (see the paragraph 'Study limitations and future directions') and/or to stochastic processes such as ecological drift and limited dispersal.

MULTI-SCALE VARIATION IN COMMUNITY COMPOSITION

We found that variation in community composition was concentrated at both fine and broad spatial scales in infralittoral fringe communities, and only at broad spatial scales in subtidal ones. Expected lower dispersal distances in infralittoral fringe than in subtidal communities may contribute to this difference, promoting finer scale variation in community composition in the infralittoral fringe. In accordance to this hypothesis, some studies have shown that genetic connectivity among populations decreased with tidal height for different organisms of rocky shores (e.g. Engel, Destombe & Valero 2004; Kelly & Palumbi 2010; Valero et al. 2011), including the two kelps L. digitata and L. hyperborea (Robuchon et al. 2014). Alternatively, our results also indicate that environmental variables were good correlates of community composition at fine scale for infralittoral fringe communities. Thus, fine-scale variability in the composition of infralittoral fringe communities may also be the result of environmental filters acting at a fine scale.

STUDY LIMITATIONS AND FUTURE DIRECTIONS

Our study permitted the identification of environmental correlates of community composition for infralittoral fringe and subtidal communities. Although this work could be refined regarding the identification of variables driving community patterns, our approach allowed us to test the correlation between observed community patterns and a set of environmental variables possibly affecting these patterns. Additional experimental approaches are needed to assess causal relationships among the environmental correlates we identified and community composition. Furthermore, some potentially important variables were not included in our framework: despite being important drivers of seaweed community structure, sea turbidity, nutrient availability and herbivores' abundances were not included because data were not available at a spatial resolution fine enough for our study. Nonetheless, other studies on these omitted variables at a coarser spatial resolution indicate strong differences among regions that might contribute to regional differences in community structure; for instance, sea turbidity is higher in St Malo Bay than in the other three regions (Gohin 2011). This limitation was partially overcome because the geophysical and hydrodynamic variables that we included in our study may influence these omitted variables. Nonetheless, future work investigating environmental drivers of kelp-dominated seaweed communities should include all pertinent variables at the appropriate scale, maybe by doing direct in situ measurements when remote sensing data are not available.

Although our outcomes are coherent with individual processes known to vary across a gradient of tidal heights, our study did not permit to identify how these local processes act together to form the patterns we observed, neither to fully understand the interplay between these local processes and biogeographic history of species. A study including all tidal heights from the high intertidal to the subtidal and covering entire biogeographical regions would be very helpful to further characterise how ecological drift, environmental filtering, competition, facilitation and dispersal interact with the biogeographical history of species to shape community patterns along the whole gradient of tidal heights. Yet, even on a set of two neighbour tidal heights, the MEMs framework permitted us to highlight that the nature of environmental correlates and the spatial scale at which they were good correlates of community composition vary across tidal heights. Therefore, this framework seems promising to broaden our understanding of community assembly across other vertical gradients, both in the sea and on land.

Author's contributions

M.R., M.V. and L.L. conceived the ideas and designed methodology; M.R. and L.L. collected the data; M.R. analysed the data; E.T. critically interpreted the first results and contributed to reorient the analyses; M.R. led the writing of the manuscript. All authors contributed critically to the drafts and gave final approval for publication.

Acknowledgements

We are extremely grateful to the numerous people that helped us during sampling including those of the 'Service Mer et Observation' of the Station Biologique de Roscoff, the divers from the Dinard Marine Lab (MNHN) and those of the 'Parc Naturel Marin d'Iroise'. We warmly thank L. Couceiro, F. Lerck, A. Boisrobert and L. Jaugeon for their precious help in specimens' identifications as well as R. Gallon for useful comments on statistical procedures. The bulk of this project was funded by an agreement with the 'Parc Naturel Marin d'Iroise' (CNRS-UPMC-PNMI, LS 64816). M.R. was supported by a PhD fellowship from the French government (Ministère de l'Enseignement Supérieur et de la Recherche). We sincerely thank two anonymous reviewers as well as the Associate Editor A. Randall Hughes for their fruitful comments which helped us to improve the manuscript.

Data accessibility

Data deposited in the Dryad Digital Repository https://doi.org/10.5061/dryad. 3s96n (Robuchon et al. 2017).

References

- Bennett, S. & Wernberg, T. (2014) Canopy facilitates seaweed recruitment on subtidal temperate reefs. *Journal of Ecology*, 102, 1462–1470.
- Borcard, D., Gillet, F. & Legendre, P. (2011) *Numerical Ecology with R.* Springer Science & Business Media, New York, NY, USA.
- Boudière, E., Maisondieu, C., Ardhuin, F., Accensi, M., Pineau-Guillou, L. & Lepesqueur, J. (2013) A suitable metocean hindcast database for the design of Marine energy converters. *International Journal of Marine Energy*, 3–4, e40–e52
- Bray, J.R. & Curtis, J.T. (1957) An ordination of the upland forest communities of southern Wisconsin. *Ecological Monographs*, 27, 325–349.
- Brodie, J.A. & Irvine, L.M. (2003) Seaweeds of the British Isles, vol. 1, Rhodo-phyta, pt. 3B, Bangiophycidae. The Natural History Museum, London, UK.

- Burrows, E. (1991) Seaweeds of the British Isles, vol. 2, Chlorophyta. British Museum (Natural History), London, UK.
- Cabioc'h, J., Floc'h, J.-Y., Le Toquin, A., Boudouresque, C.F., Meinesz, A. & Verlaque, M. (2006) Guide des algues des mers d'Europe. Delachaux et Niestlé, Paris, France.
- Chase, J.M. (2007) Drought mediates the importance of stochastic community assembly. Proceedings of the National Academy of Sciences of the United States of America, 104, 17430–17434.
- Chase, J.M. & Myers, J.A. (2011) Disentangling the importance of ecological niches from stochastic processes across scales. *Philosophical Transactions of* the Royal Society of London B Biological Sciences, 366, 2351–2363.
- Christie, H., Jorgensen, N.M., Norderhaug, K.M. & Waage-Nielsen, E. (2003) Species distribution and habitat exploitation of fauna associated with kelp (*Laminaria hyperborea*) along the Norwegian coast. *Journal of the Marine Biological Association of the UK*, 83, 687–699.
- Cottenie, K. (2005) Integrating environmental and spatial processes in ecological community dynamics. *Ecology Letters*, 8, 1175–1182.
- Cowen, R.K. (2000) Connectivity of marine populations: open or closed? Science, 287, 857–859.
- Crist, T.O., Veech, J.A., Gering, J.C. & Summerville, K.S. (2003) Partitioning species diversity across landscapes and regions: a hierarchical analysis of α , β , and γ diversity. *The American Naturalist*, **162**, 734–743.
- Dayton, P.K. (1985) Ecology of kelp communities. Annual Review of Ecology and Systematics, 16, 215–245.
- Derrien-Courtel, S., Le Gal, A. & Grall, J. (2013) Regional-scale analysis of subtidal rocky shore community. Helgoland Marine Research, 67, 697–712.
- Dixon, P.S. & Irvine, L.M. (1977) Seaweeds of the British Isles, vol. 1, Rhodophyta, pt. 1, Introduction, Nemaliales, Gigartinales. British Museum (Natural History), London, UK.
- Dray, S., Legendre, P. & Peres-Neto, P.R. (2006) Spatial modelling: a comprehensive framework for principal coordinate analysis of neighbour matrices (PCNM). *Ecological Modelling*, 196, 483–493.
- Dray, S., Pélissier, R., Couteron, P. et al. (2012) Community ecology in the age of multivariate multiscale spatial analysis. *Ecological Monographs*, 82, 257–275.
- Engel, C.R., Destombe, C. & Valero, M. (2004) Mating system and gene flow in the red seaweed *Gracilaria gracilis*: effect of haploid-diploid life history and intertidal rocky shore landscape on finescale genetic structure. *Heredity*, 92, 289–298.
- Fletcher, R.L. (1987) Seaweeds of the British Isles, vol. 3, Fucophyceae (Phaeophyceae) pt. 1. British Museum (Natural History), London, UK.
- Fraschetti, S., Terlizzi, A. & Benedetti-Cecchi, L. (2005) Patterns of distribution of marine assemblages from rocky shores: evidence of relevant scales of variation. *Marine Ecology Progress Series*, 296, 13–29.
- Gallon, R.K., Robuchon, M., Leroy, B., Le Gall, L., Valero, M. & Feunteun, E. (2014) Twenty years of observed and predicted changes in subtidal red seaweed assemblages along a biogeographical transition zone: inferring potential causes from environmental data. *Journal of Biogeography*, 41, 2293–2306.
- Gohin, F. (2011) Annual cycles of chlorophyll-a, non-algal suspended particulate matter, and turbidity observed from space and in-situ in coastal waters. Ocean Science, 7, 705–732.
- Graham, M.H. (2004) Effects of local deforestation on the diversity and structure of southern California giant kelp forest food webs. *Ecosystems*, 7, 341–357.
- Grime, J.P. (1973) Competitive exclusion in herbaceous vegetation. *Nature*, 242, 344–347.
- Hawkins, S. & Harkin, E. (1985) Preliminary canopy removal experiments in algal dominated communities low on the shore and in the shallow subtidal on the Isle of Man. *Botanica Marina*, 28, 223–230.
- Hubbell, S.P. (2001) The Unified Neutral Theory of Biodiversity and Biogeography. Princeton University Press, Princeton, NJ, USA.
- Hutchinson, G.E. (1957) Concluding remarks. Cold Spring Harbor Symposia on Quantitative Biology. Reprinted in: Classics in Theoretical Biology. Bulletin of Mathematical Biology, 53, 193–213.
- Irvine, L.M. (1983) Seaweeds of the British Isles, vol. 1, Rhodophyta, pt. 2A, Cryptonemiales (sensu stricto), Palmariales, Rhodymeniales. British Museum (Natural History), London, UK.
- Irvine, L.M., Chamberlain, Y.M. & Maggs, C.A. (1994) Seaweeds of the British Isles, vol. 1, Rhodophyta, pt. 2B, Corallinales, Hildenbrandiales. The Natural History Museum, London, UK.
- Kain, J.M. (1979) A view of the genus Laminaria. Oceanography and Marine Biology: An Annual Review, 17, 101–161.
- Keith, S.A., Kerswell, A.P. & Connolly, S.R. (2014) Global diversity of marine macroalgae: environmental conditions explain less variation in the tropics. *Global Ecology and Biogeography*, 23, 517–529.

- Kelly, R.P. & Palumbi, S.R. (2010) Genetic structure among 50 species of the northeastern Pacific rocky intertidal community, PLoS ONE, 5, e8594.
- Kerswell, A.P. (2006) Global biodiversity patterns of benthic marine algae. Ecology, 87, 2479-2488.
- Kindt, R. & Coe, R. (2005) Tree Diversity Analysis: A Manual and Software for Common Statistical Methods for Ecological and Biodiversity Studies. World Agroforestry Centre, Nairobi, Kenya.
- Kinlan, B.P. & Gaines, S.D. (2003) Propagule dispersal in marine and terrestrial environments: a community perspective. Ecology, 84, 2007–2020.
- Kuhn, M. Contributions from Jed Wing, S.W., Andre Williams, Chris Keefer, Allan Engelhardt, Tony Cooper, Zachary Mayer, Brenton Kenkel, the R Core Team, Michael Benesty, Reynald Lescarbeau, Andrew Ziem, Luca Scrucca, Yuan Tang and Can Candan (2015) caret: Classification and Regression Training. R package version 6.0-68. Available at: https://CRAN.R-project. org/package=caret (accessed 24 April 2017).
- Le Fèvre, J. (1987) Aspects of the biology of frontal systems. Advances in Marine Biology, 23, 163-299.
- Leclerc, J.-C., Riera, P., Laurans, M., Leroux, C., Lévêque, L. & Davoult, D. (2015) Community, trophic structure and functioning in two contrasting Laminaria hyperborea forests, Estuarine, Coastal and Shelf Science, 152, 11-22.
- Leibold, M.A., Holyoak, M., Mouquet, N. et al. (2004) The metacommunity concept: a framework for multi-scale community ecology. Ecology Letters, 7, 601-613.
- Leibold, M.A. & McPeek, M.A. (2006) Coexistence of the niche and neutral perspectives in community ecology. Ecology, 87, 1399-1410.
- Leliaert, F., Anderson, R.J., Bolton, J.J. & Coppejans, E. (2000) Subtidal understorey algal community structure in kelp beds around the Cape Peninsula (Western Cape, South Africa). Botanica Marina, 43, 359-366.
- Logue, J.B., Mouquet, N., Peter, H. & Hillebrand, H. (2011) Empirical approaches to metacommunities: a review and comparison with theory. Trends in Ecology & Evolution, 26, 482-491.
- Maggs, C.A. & Hommersand, M.H. (1993) Seaweeds of the British Isles, vol. 1, Rhodophyta, pt. 3A, Ceramiales. The Natural History Museum, London,
- Mann, K.H. (1973) Seaweeds: their productivity and strategy for growth. Science, 182, 975-981.
- McIntire, E.J. & Fajardo, A. (2009) Beyond description: the active and effective way to infer processes from spatial patterns. Ecology, 90, 46-56.
- Miller, R.J. & Etter, R.J. (2008) Shading facilitates sessile invertebrate dominance in the rocky subtidal Gulf of Maine. Ecology, 89, 452-462.
- Munoz, F. (2009) Distance-based eigenvector maps (DBEM) to analyse metapopulation structure with irregular sampling. Ecological Modelling, 220,
- Myers, J.A. & Harms, K.E. (2011) Seed arrival and ecological filters interact to assemble high-diversity plant communities, Ecology, 92, 676–686.
- Norton, T.A. (1992) Dispersal by macroalgae. British Phycological Journal, 27,
- Oksanen, J., Blanchet, G.F., Kindt, R. et al. (2015) vegan: Community Ecology Package. R package version 2.3-5. Available at: https://CRAN.R-project.org/ package=vegan (accessed 24 April 2017).
- Ollier, S., Couteron, P. & Chessel, D. (2006) Orthonormal transform to decompose the variance of a life-history trait across a phylogenetic tree. Biometrics, **62**, 471-477.
- Oppliger, L.V., von Dassow, P., Bouchemousse, S., Robuchon, M., Valero, M., Correa, J.A., Mauger, S. & Destombe, C. (2014) Alteration of sexual reproduction and genetic diversity in the kelp species Laminaria digitata at the southern limit of its range. PLoS ONE, 9, e102518.
- Pehlke, C. & Bartsch, I. (2008) Changes in depth distribution and biomass of sublittoral seaweeds at Helgoland (North Sea) between 1970 and 2005. Climate Research, 37, 135-147.
- Platt, T. & Denman, K.L. (1975) Spectral analysis in ecology. Annual Review of Ecology and Systematics, 6, 189-210.
- R Core Team (2015) R: A Language and Environment for Statistical Computing. R Foundation for Statistical Computing, Vienna, Austria.
- Raffaelli, D. & Hawkins, S. (1996) Intertidal Ecology. Chapman & Hall, London, UK.
- Rao, C.R. (1995) A review of canonical coordinates and an alternative to correspondence analysis using Hellinger distance. Questiió: Quaderns d'Estadística, Sistemes, Informatica i Investigació Operativa, 19, 23-63.
- Robuchon, M., Le Gall, L., Mauger, S. & Valero, M. (2014) Contrasting genetic diversity patterns in two sister kelp species co-distributed along the coast of Brittany, France. Molecular Ecology, 23, 2669-2685.

- Robuchon, M., Valero, M., Gey, D. & Le Gall, L. (2015) How does molecularassisted identification affect our estimation of α , β and γ biodiversity? An example from understory red seaweeds (Rhodophyta) of Laminaria kelp forests in Brittany, France. Genetica, 143, 207-223.
- Robuchon, M., Valero, M., Thiébaut, E. & Le Gall, L. (2017) Data from: Multi-scale drivers of community diversity and composition across tidal heights: an example on temperate seaweed communities. Dryad Digital Repository, https://doi.org/10.5061/dryad.3s96n
- Salomon, J.-C. & Breton, M. (1993) An atlas of long-term currents in the Channel. Oceanologica Acta, 16, 439-448.
- Sbrocco, E.J. & Barber, P.H. (2013) MARSPEC: ocean climate layers for marine spatial ecology: Ecological Archives E094-086. Ecology, 94, 979-979.
- Shannon, C.E. (1948) A mathematical theory of communication. Bell System Technical Journal, 27, 379-423.
- Smale, D.A., Kendrick, G.A. & Wernberg, T. (2011) Subtidal macroalgal richness, diversity and turnover, at multiple spatial scales, along the southwestern Australian coastline. Estuarine, Coastal and Shelf Science, 91, 224-231.
- Spalding, M.D., Fox, H.E., Allen, G.R. et al. (2007) Marine ecoregions of the world: a bioregionalization of coastal and shelf areas. BioScience, 57, 573-583.
- Steneck, R.S., Graham, M.H., Bourque, B.J., Corbett, D., Erlandson, J.M., Estes, J.A. & Tegner, M.J. (2002) Kelp forest ecosystems: biodiversity, stability, resilience and future. Environmental Conservation, 29, 436-459.
- Tilman, D. (1982) Resource Competition and Community Structure. Princeton University Press, Princeton, NJ, USA.
- Valero, M., Destombe, C., Mauger, S., Ribout, C., Engel, C.R., Daguin-Thiebaut, C. & Tellier, F. (2011) Using genetic tools for sustainable management of kelps: a literature review and the example of Laminaria digitata, Cahiers de Biologie Marine, 52, 467-483.
- Vanderklift, M.A., Lavery, P.S. & Waddington, K.I. (2009) Intensity of herbivory on kelp by fish and sea urchins differs between inshore and offshore reefs. Marine Ecology Progress Series, 376, 203-211.
- Wernberg, T., Bennett, S., Babcock, R.C. et al. (2016) Climate-driven regime shift of a temperate marine ecosystem. Science, 353, 169-172.
- Wernberg, T. & Connell, S.D. (2008) Physical disturbance and subtidal habitat structure on open rocky coasts: effects of wave exposure, extent and intensity. Journal of Sea Research, 59, 237-248.
- Wernberg, T., Kendrick, G.A. & Phillips, J.C. (2003) Regional differences in kelp-associated algal assemblages on temperate limestone reefs in south-western Australia. Diversity and Distributions, 9, 427-441.
- Wernberg, T., Kendrick, G.A. & Toohey, B.D. (2005) Modification of the physical environment by an Ecklonia radiata (Laminariales) canopy and implications for associated foliose algae. Aquatic Ecology, 39, 419-430.

Received 4 December 2016; accepted 29 March 2017 Handling Editor: A. Randall Hughes

Supporting Information

Details of electronic Supporting Information are provided below.

- Appendix S1. Taxonomic and geographic information on seaweed specimens collected for the study.
- Appendix S2. Matrix of pairwise geographic distances among the sites of the study.
- Appendix S3. Site-by-environment matrix containing explanatory environmental variables for infralittoral fringe communities.
- Appendix S4. Site-by-environment matrix containing explanatory environmental variables for subtidal communities.
- Appendix S5. R script to reproduce the analyses using data of Appendices S1-S4, Supporting Information.