Trophic opportunism of central Amazon floodplain fish

J. M. MORTILLARO*, M. POUILLY*, M. WACH*, C. E. C. FREITAS[†], G. ABRIL^{‡,§} AND T. MEZIANE* *Biology of Aquatic Organisms and Ecosystems (BOREA), Université Pierre et Marie Curie, Sorbonne Universités, Muséum National d'Histoire Naturelle, Université Caen Basse Normandie, UMR-CNRS 7208, IRD 207, CP 53, Paris, Cedex 05, France [†]Departamento de Ciências Pesqueiras, Faculdade de Ciências Agrárias, Universidade Federal do Amazonas, Manaus, AM, Brazil [‡]Laboratoire Environnement et Paléoenvironnements Océanique (EPOC) UMR-CNRS 5805, Université de Bordeaux, Pessac, France [§]Programa de Geoquímica, Universidade Federal Fluminense, Niterói, RJ, Brazil

SUMMARY

1. The food web of the central Amazon basin displays one of the largest discrepancies in food source utilisation versus availability for consumers. While C_4 macrophytes dominate the primary producing biomass in floodplains, the food web is dominated by the use of C_3 carbon sources. Amazon fish species have wide-ranging diets and show feeding flexibility in response to spatial and temporal patterns in food source availability. Fish are therefore expected to use a range of available resources. Fatty acids and stable carbon and nitrogen isotopic compositions were studied to investigate their trophic behaviour and provide a better assessment of food source utilisation by nine fish species from two Amazon floodplains.

2. Although nitrogen stable isotopes positioned detritivores at the base of the food chain, their large proportions of polyunsaturated fatty acids (FAs) suggest selection of a high-quality food source, such as microalgae. Omnivores and insectivores displayed a wide range of food source utilisation, indicating opportunist feeding behaviour. Piscivores displayed the most ¹⁵N-enriched signature (up to 10.6%), concomitant with their predator position in the studied food web.

3. C_3 carbon was the dominant source for most species, which is a finding consistent with current classical knowledge. However, consumption of C_4 macrophytes was detected for the herbivore *Schizodon fasciatus*, in line with the spatial availability of this food source in the floodplain system. Large amounts of polyunsaturated and essential FAs in these C_4 macrophytes indicate their high nutritive value. Our results suggest that this food source is rather neglected by fish due to low digestibility, with the exception of the herbivore that may possess modification of its digestive system that allows the use of C_4 plants.

Keywords: Amazon, fatty acids, fish, floodplains, stables isotopes

Introduction

Primary production is considered a main structuring factor of food webs in aquatic systems. Both terrestrial and aquatic sources of production are involved. However, the relationship between source production and use by consumers is not always clear. Evaluation of organic matter (OM) sources and transfer pathways through the food web is essential to understand ecosystem functioning. Fish production results from a series of transformations of organic carbon through the food web, beginning with carbon dioxide fixation by primary producers, such as plants and phytoplankton (Forsberg *et al.*, 1993).

In the Amazon system, primary production is dominated by aquatic macrophytes (52–64%), followed by flooded forest trees (28–32%), periphyton (5–8%) and phytoplankton (2–8%) (Junk, 1985; Melack *et al.*, 1999). Among macrophytes, the perennial grasses *Echinochloa polystachya*, *Paspalum fasciculatum* and *Paspalum repens* can reach cumulative biomasses of 160 t ha⁻¹ (Piedade, Junk & Long, 1991; Junk & Piedade, 1993). These species, which convert carbon dioxide into

Correspondence: Jean-Michel Mortillaro, Biology of Aquatic Organisms and Ecosystems (BOREA), Université Pierre et Marie Curie, Sorbonne Universités, Muséum National d'Histoire Naturelle, Université Caen Basse Normandie, UMR-CNRS 7208, IRD 207, CP 53, Paris Cedex 05, France. E-mail: mortillaro@mnhn.fr

biomass through a C₄ pathway, can form floating meadows (Junk & Howard-Williams, 1984) in which macrophytes using the C₃ photosynthetic pathway, such as Eichornia sp., Pistia stratiotes and Salvinia auriculata, are also found. However, the total biomass of C3 macrophytes is lower than that of C_4 plants (3 to 15 t ha⁻¹; Furch & Junk, 1992). If ecosystem functioning involved consumption of the most available source of carbon, a dominant contribution of C₄ aquatic macrophytes to the organic matter of upper trophic levels would be expected. However, various studies characterising the food web of Amazon floodplains have reported a dominance of C₃ carbon sources, including macrophytes, microalgae and trees (Araujo-Lima et al., 1986; Forsberg et al., 1993; Oliveira et al., 2006). Moreover, recent studies suggest a stronger than expected contribution of phytoplankton to particulate organic matter (POM) primary production in the central Amazon Várzea (Mortillaro et al., 2011). There remains, therefore, uncertainties about the fate of the C₄ carbon in the Amazon ecosystem, which might relate to neglect by consumers, although some studies have suggested that bacterial degradation could be the main driver (Quay et al., 1992; Marshall, Forsberg & Thome-Souza, 2008; J. M. Mortillaro unpubl. data).

In the central Amazon basin, many studies use carbon stable isotopes (SI, δ^{13} C) to determine fish autotrophic food sources (Araujo-Lima et al., 1986; Forsberg et al., 1993; Benedito-Cecilio & Araujo-Lima, 2002; Oliveira et al., 2006; Marshall et al., 2008). Carbon isotopic ratios can be useful in distinguishing terrestrial from aquatic producers. Isotopic signatures are not always easy to distinguish in freshwater ecosystems, however, particularly when organisms receive OM contributions from various origins, such as phytoplankton or aquatic, semiaquatic and terrestrial C₃ primary producers (Hedges et al., 1986; Townsend-Small, McClain & Brandes, 2005). Thus, studies in freshwater ecosystems often use both SI and fatty acids (FAs; Canuel et al., 1995; Alfaro et al., 2006). FAs are carbon-rich compounds ubiquitous in all organisms and relatively easy to metabolise and store (Alfaro et al., 2006). Some FAs are only biosynthesised by specific organisms and transferred without modification from primary producers towards higher trophic levels (Parrish et al., 2000; Dalsgaard et al., 2003). Thus, according to the tissue analysed, FAs may provide taxonomic information (Kwetegyeka et al., 2011) or may help characterise food sources and quality within trophic webs (Ahlgren *et al.*, 1994; Koussoroplis *et al.*, 2011).

In the Amazon basin, food source production is modulated by the hydrological cycle and local inputs (e.g. water, OM, nutrients) result from landscape variations (i.e. upland basin area, maximum floodable area and vegetation cover; Junk, 1997). Also, due to the large variability and ephemeral character of habitats and resources of riverine ecosystems, fish species are generally considered as mostly opportunistic in their feeding habits (Lowe-McConnell, 1987; Araujo-Lima, Agostinho & Fabre, 1995; Mérona & Mérona, 2004). Within this context, the objectives of this study were to provide an assessment of food sources for nine fish species, as well as spatial variations in their utilisation. FAs and SI (δ^{13} C, δ^{15} N) compositions of fish were compared in order to depict the trophic structure and function of two Amazon floodplains and to evaluate the fate of the C₄ carbon in upper trophic levels.

Methods

Study sites

Samples were collected in two central Amazon floodplains, Cabaliana and Janauacá, adjacent to the Solimões River, near the confluence of the Solimões and Negro rivers, upstream of Manaus city (Fig. 1a). These floodplains are mainly supplied by white water from the Solimões River, which is characterised by high concentrations of sedimentary material and dissolved and POM (Sioli, 1984). A large diversity of vegetation typical of the central Amazon floodplain, including flooded forest and shrubs, floating macrophytes and phytoplankton in open waters, is present in these floodplains (Hess et al., 2003). Two sampling programmes occurred during a similar hydrological period: the beginning of the low water period (c. 4 months after the flood peak) in October 2009 and during the period of falling water level (c. 3 months after the flood peak) in August 2010 (Fig. 1b, Table 1).

Sampling

Fish were captured using 2.5 m by 25 m gillnets with a wide range of mesh size (20, 25, 30, 35, 40 and 50 mm). Major habitats of each floodplain were sampled, including open waters, non-vegetated banks, floating meadows and flooded forests. Gillnets were left fishing perpendicular to the bank for 2 h in the evening (17:00–19:00 hours) and for 2 h in the morning (5:00–7:00 hours). Captured fish were immediately identified to the species level, and non-selected living fish were returned to the water. For selected individuals, fish were sacrificed rapidly after capture to minimise suffering



Fig. 1 (a) Sampling sites on the Solimões River upstream of Manaus city (Brazil). Dashed lines delineate the Cabaliana and Janauacá floodplains (Abril *et al.*, 2014). (b) Hydrograph at the Manaus gauging station of the Amazon River. Solid line is for the 2009 hydrograph and dashed line for 2010, with bold areas highlighting the sampling periods.

Table 1 Surface and proportional area of floodplain vegetation inCabaliana and Janauacá during October 2009 and August 2010(Abril *et al.*, 2014)

Surface in km ²	Cabaliana (%)	Janauacá (%)
Low water, October 2009		
Open waters	139 (50.4)	48 (92.3)
Flooded forest	111 (40.2)	1 (1.9)
Floating vegetation	26 (9.4)	3 (5.8)
Total floodplain	276	52
Falling water, August-Sep	tember 2010	
Open waters	210 (29.2)	62 (75.6)
Flooded forest	464 (64.5)	13 (15.9)
Floating vegetation	45 (6.3)	7 (8.5)
Total floodplain	719	82

(by stunning or freezing). Fish were photographed and measured (standard length), and a 3 to 5 g of anterodorsal muscle was sampled.

Nine fish species were selected, based on the frequency of capture. According to the literature, five dietary groups are found in the nine selected species (Table 2): detritivores, herbivores, insectivores, omnivores and piscivores. The herbivores were separated into three subgroups: consuming mainly algae, mainly macrophytes or mainly fruits. Trophic groups used in this study represent general tendencies and not strict diets based exclusively on one food source. Differences in diet are commonly observed according to developmental stage and fish size (Merigoux & Ponton, 1998). Consequently, for each species, only large adult individuals were selected.

Potential food sources were also sampled, including POM (Whatman GF/F filters burned at 450 °C for 12 h), surface sediments, plankton net contents (63 μ m mesh size) and aquatic macrophytes.

Samples analysis

Fish and food source samples were immediately frozen and freeze-dried after collection for FA and SI analyses. FA composition of samples was obtained according to a modification of the method of Bligh & Dyer (1959) previously described in Mortillaro et al. (2011). FA analyses were performed in the Resaqua Laboratory (MNHN, Paris), using gas chromatography coupled either to a flame ionisation detector (Varian GC 3800CP, Varian, Inc., Walnut Creek, California, USA) or a mass spectrometer (Varian 200MS, Varian, Inc., Walnut Creek, California, USA). Carbon $(\delta^{13}C, {}^{13}C/{}^{12}C)$ and nitrogen $(\delta^{15}N,$ ¹⁵N/¹⁴N) isotopic ratios were calculated in parts per thousand (%) in relation to an international standard (Vienna PeeDee Belemnite for $\delta^{13}C$ and atmospheric nitrogen for δ^{15} N; Peterson & Fry, 1987). Isotopic results were determined at the Davis Stable Isotope Facility, U.S.A., using a Europe Hydra 20/20 spectrometer. Nylon (G-11, n = 59, SD = 0.058 and 0.131 for δ^{13} C and δ^{15} N, respectively), enriched glutamic acid (G-12, n = 8, SD = 0.212 and 0.297 for δ^{13} C and δ^{15} N, respectively), bovine liver (G-13, n = 4, SD = 0.164 and 0.07 for δ^{13} C and δ^{15} N, respectively), peach leaves (G-7, n = 32, SD = 0.119 and 0.131 for δ^{13} C and δ^{15} N, respectively) and non-enriched glutamic acid (G-9, n = 2, SD = 0.012 and 0.3 for δ^{13} C and δ^{15} N, respectively) were used as internal isotope standards.

Data analysis

The FA composition and isotopic signatures of fish species were compared by species and trophic groups to highlight spatial variations (differences between the two floodplains).

All identified FAs were used in these analyses without transformation, and individual FAs are represented by their relative proportions. Dissimilarity in FA composition (%) between two individuals was evaluated using the Bray–Curtis coefficient. A dissimilarity matrix between each pair of individuals was created and used as a basis to evaluate the statistical differences in FA

4 J. M. Mortillaro et al.

Table 2 (Classification	and trophic	groups of th	e nine fish	species sampled
-----------	----------------	-------------	--------------	-------------	-----------------

Order	Family	Species	Common name (Brazil)	Ν	Size range (mm)	Trophic group	Reference
Silur	Loricariidae	Pterygoplichthys multiradiatus	Bodó	11	225–363	Detritivore	Mérona & Mérona (2004)
Chara	Prochilodontidae	Semaprochilodus insignis	Jaraqui	9	150–180	Detritivore	Benedito-Cecilio & Araujo- Lima (2002)
Chara	Hemiodontidae	Hemiodus microlepis	Orana	7	160–220	Herbivore (algae)	Silva et al. (2008)
Chara	Characidae– Serrasalminae	Mylossoma aureum	Pacu	12	122–180	Herbivore (fruits)	Mérona & Mérona (2004)
Chara	Anostomidae	Schizodon fasciatus	Aracu	20	190–283	Herbivore (roots)	Soares <i>et al.</i> (1986)
Chara	Characidae– Triporthinae	Triportheus angulatus	Sardinha	8	155–180	Insectivore	Yamamoto, Soares & Freitas (2004)
Chara	Anostomidae	Leporinus friderici	Aracu cabeça gorda	5	155–230	Omnivore	Goulding (1980)
Chara	Characidae- Serrasalminae	Pygocentrus nattereri	Piranha	11	130–145	Piscivore	Mérona & Mérona (2004)
Perci	Cichlidae	Cichla monoculus	Tucunare	8	155–320	Piscivore	Oliveira et al. (2006)

Silur, Siluriformes; Chara, Characiformes; Perci, Perciformes.

proportions. Distances between individuals were graphically represented using nMDS (non-metric multidimensional scaling; Zuur, Ieno & Smith, 2007). Between-group differences were analysed with similarity tests (ANOSIM, with 5000 permutations; Oksanen *et al.*, 2010). When there were significant differences, a similarity of percentage analysis (SIMPER; Oksanen *et al.*, 2010) was used to determine the relative contribution of each FA to between-groups differences.

The differences in δ^{13} C and δ^{15} N isotopic signatures between species (n = 9) and floodplains (n = 2) were tested using a nonparametric Kruskal–Wallis test (KW). A nonparametric Mann–Whitney–Wilcoxon (MWW) *post hoc* test was then used to identify differences between independent groups, after correction following a Benjamini and Hochberg (BH) correction procedure (Benjamini & Hochberg, 1995) for multiple comparisons. All statistical analyses were implemented in the R programming environment (R Development Core Team 2010, package Vegan; Oksanen *et al.*, 2010), with the probability α set at 0.05.

Results

Characterisation of trophic variability

Forty-five FAs were identified in the fish samples (Fig. 2), among which eight represented a cumulative contribution of 68.4 to 85.1% of the total FAs of each species. Fish species displayed significant differences in

their global FA composition (ANOSIM, R = 0.55, p < 0.001). Similarity in the FA composition within trophic groups was higher than 78% (Table 3), whereas intergroup similarities ranged from 59.7 to 81.3% (Table 4). The FA composition of the frugivore M. aureum was dominated by 18:1w9, 18:2w6 and 20:0 and had intergroup similarity (from 59.7 to 73.9%, Table 4). Differences among the other diet groups were lower, and several species from different groups could not be differentiated; for example, both herbivores H. microlepis and S. fasciatus (R = 0.12, p = 0.13) and also H. microlepis were indistinguishable from the piscivore P. nattereri (R = 0.07, p = 0.21). Moreover, no significant differences were found between S. fasciatus and P. nattereri (R = -0.03, p = 0.56), or between the omnivore L. frideri*ci* and the insectivore *T. angulatus* (R = 0.07, p = 0.24). Global patterns of FA composition were observed (Fig. 2, Table 3) with a stronger proportion of $20:4\omega 6$ and 22:6w3 in piscivores and detritivores; a stronger proportion of 18:109 and 18:206 in frugivores, insectivores and omnivores; and a stronger proportion of 16:1w7 in detritivores and algivores. Herbivore feeding predominantly on roots (S. fasciatus) contained a stronger proportion of 18:303 and 20:406. The FA composition of the sampled food sources is summarised in Tables S1 and S2.

Carbon isotope signatures of the sampled food sources ranged from -41 to -29%, with the exception of C₄ macrophytes, which were ¹³C enriched ($-13.1 \pm 5.4\%$). Food source nitrogen isotope signatures ranged between



Fig. 2 Non-metric MDS of individual fatty acid (FA) proportion (%) in fish with different diets. Each point represents a sample collected in Janauacá (Jan., filled symbol) or Cabaliana (Cab., open symbol), with piscivores in red, insectivores in black, omnivores in orange, detritivores in brown and herbivores (fruits – pink; algae – green; roots – blue). FAs in bold displayed an average proportion greater than 1% of total FAs.

Table 3 Summary of SIMPER procedure for mean fatty acid compositions and intragroup similarity for each trophic group

Trophic group (intragroup similarity)	16:0	18:0	16:1ω7	18:1ω9	18:2ω6	18:3ω3	20:4ω6	22:6w3	Σ (%)
Detritivores (85.6%)	22.3	10.5	5.9	9.9	4.4	2.2	8.1	5.1	68.4
Herbivore roots (82.4%)	27.6	8.8	3.1	15.3	5.3	3.7	7.2	3.7	74.7
Herbivore fruits (78.7%)	28.0	14.3	2.0	25.9	11.3	1.0	1.4	1.2	85.1
Herbivore algae (82.9%)	28.0	9.4	5.3	14.1	2.5	2.9	6.1	5.0	73.3
Insectivores (80.1%)	23.9	9.7	4.6	25.9	6.9	2.1	4.9	4.8	82.8
Omnivores (84.8%)	28.3	10.4	3.2	25.1	4.0	1.0	5.8	5.3	83.1
Piscivores (82.7%)	22.9	12.1	2.5	10.9	3.8	2.0	9.4	10.0	73.6

0.5 and 6% (Fig. 3). Fish isotope signatures ranged from -37 to -21% for δ^{13} C and 5 to 11% for δ^{15} N. Detritivores and frugivores were the most ¹⁵N-depleted species and piscivores the most ¹⁵N enriched. Additionally, isotopic signatures revealed interspecific differences. The carbon isotope signature of M. aureum differed significantly from all other species (MWW, P < 0.05) except S. fasciatus. However, the nitrogen isotope signature of M. aureum was not significantly different from the omnivore and detritivores. The $\delta^{15}N$ signature from two detritivores, S. insignis and P. multiradiatus, was significantly different from all other species (except L. friderici). However, the carbon isotope signatures of both detritivores were not significantly different from other species (p > 0.05), with the exception of *H. microlepis* (p = 0.04, P. multiradiatus and H. microlepis). The algivore, H. microlepis, also displayed significant differ-

© 2015 John Wiley & Sons Ltd, Freshwater Biology, doi: 10.1111/fwb.12598

ences in δ^{13} C from other species (except *C. monoculus*), but only with piscivores and insectivores for δ^{15} N. All remaining species (*C. monoculus*, *P. nattereri*, *T. angulatus*, *L. friderici* and *S. fasciatus*) displayed significant differences among them for δ^{15} N, with the exception of *C. monoculus* and *L. friderici* (p > 0.05) as well as *L. friderici* and *S. fasciatus* (p > 0.05). However, no difference in carbon isotopic signature was found for these species.

Spatial variability

Significant differences were observed in FA composition between the two floodplains (ANOSIM, R = 0.07, p = 0.03). The large intragroup variability (R close to 0) may be a result of differences between species. A spatial difference between the two sampled floodplains was recorded for *S. fasciatus* (R = 0.28, p = 0.01, Figs 2 & 4)

6 J. M. Mortillaro et al.

Table 4 Summary of SIMPER procedure for intergroup similarity for each trophic group

Intergroup similarity (%)	Detritivores	Herbivore-roots	Herbivore-fruits	Herbivore-algae	Insectivores	Omnivores
Herbivore-roots	78.9					
Herbivore-fruits	59.7	66.7				
Herbivore-algae	79.8	81.3	64.8			
Insectivores	73.4	76.4	73	75.8		
Omnivores	73.7	79.6	73.9	79	80.4	
Piscivores	80.4	77.1	60.7	76.9	72.4	74.5



Fig. 3 δ^{13} C and δ^{15} N (*n*) biplot of sampled fish and potential food sources. For fish, each point represents a sample collected in Janauacá (filled symbol) or Cabaliana (open symbol), with piscivores in red, insectivores in black, omnivores in orange, detritivores in brown and herbivores (fruits – pink; algae – green; roots – blue). For sources, each point highlights the mean value and standard deviation from both floodplains, with leaves in green, roots in black, particulate organic matter and sediments in blue.

and for piscivores (R = 0.32, p = 0.02, sampled in 2010). In contrast, no spatial difference was recorded for any of the other species: detritivores (R = 0.04, p = 0.23), algivores (R = 0.43, p = 0.06), frugivores (R = 0.03, p = 0.29) and insectivores (R = -0.14, p = 0.70).

The isotopic composition of fish between the two floodplains also differed significantly (Fig. 3). Spatial differences in δ^{13} C and δ^{15} N composition were observed for *S. fasciatus*, *P. nattereri* and *M. aureum*. In addition, *M. aureum* was ¹⁵N depleted in Cabaliana in comparison with Janauacá (MWW, p = 0.02), and samples of *S. fasciatus* from Cabaliana were ¹⁵N (p = 0.002) and ¹³C (p = 0.0004, Figs 3 & 4) enriched in comparison with those from Janauacá. Furthermore, samples of *P. nattereri* from Cabaliana were ¹³C depleted in comparison with those from Janauacá (p = 0.04).

Discussion

In the nine studied species, 45 FAs were identified, a large proportion being mono- and polyunsaturated.

Mono- and polyunsaturated FAs (with the exception of the w3 series, e.g. 20:5w3 and 22:6w3) are primarily catabolised during starvation (Koven, Kissil & Tandler, 1989; Izquierdo, 1996; Rainuzzo, Reitan & Olsen, 1997), suggesting that fish in our study were not starved. Of the eight major FAs present in fish samples, six were essential fatty acids (EFAs) or EFA precursors, which cannot be synthesised by fish but must be obtained through feeding (Holman, 1998; Arts, Ackman & Holub, 2001). Additional FAs each represented less than 1% of the total, but their global diversity is in accord with other Amazon fish assemblages (Inhamuns & Franco, 2001; Inhamuns, Franco & Batista, 2009). Our aquatic ecosystem can thus be considered to be in an equilibrium state, without major limiting food resources (Wetzel, 2001). Isotopic results also correspond to the classical notion of food-web organisation. In agreement with other Amazon studies, $\delta^{15}N$ enrichment was observed among the successive trophic levels and $\delta^{13}C$ indicated a predominance of C₃ food sources (Forsberg et al., 1993; Oliveira et al., 2006).



total fatty acid (FA) concentration (mg g^{-1}) and FA marker proportion in *S. fasciatus* muscles sampled in Cabaliana and Janauacá.

Fig. 4 Summary of δ^{13} C composition,

Food source utilisation

The predominance of detritus in gut contents means P. multiradiatus is considered a detritivore (Mérona & Mérona, 2004). This species displayed one of the most ¹⁵N-depleted signatures (6.7 \pm 0.8%), positioning it at the base of the food chain. Similarly, S. insignis preferentially consumes detritus aggregates of algae and microorganisms (Goulding, Carvalho & Ferreira, 1988); its $\delta^{15}N$ signature closely resembles that of *P. multiradiatus* $(6.2 \pm 0.8\%)$. These two species usually display an affinity for C₄ macrophytes (P. multiradiatus; Araujo-Lima et al., 1986; Forsberg et al., 1993) and for phytoplankton (S. insignis; Araujo-Lima et al., 1986). Our isotopic composition of POM and S. insignis confirms a preference for phytoplankton. However, ¹³C-depleted signatures $(-31.8 \pm 2.3\%$ and $-33.0 \pm 3.0\%$ for P. multiradiatus and S. insignis, respectively) suggest that both species avoid C₄ macrophytes. Such ¹³C-depleted signatures contradict results from Forsberg et al. (1993; -23.8% for P. multiradiatus), where the dietary contribution of carbon from C₄ macrophytes was estimated to range from 10.4 to 41.5%. Instead, our results suggest a greater contribution of microalgae to the diet of these detritivores. This finding is further supported by the greater proportion in these species of 16:1w7, 18:1w9 and 22:6w3, FAs

associated with cyanobacteria and microalgae in freshwater ecosystems (Napolitano, 1999; Mortillaro et al., 2011). These species do not have anatomical adaptations to filter-feed on phytoplankton (Goulding et al., 1988). Forsberg et al. (1993) assumed that OM consumed by these fish is either derived from (1) the phytoplankton layer decomposing on sediments or (2) from the selective assimilation of phytoplankton in detritus. The former hypothesis is unlikely, due to slight differences between POM and sediment isotopic compositions. The second hypothesis appears more plausible and is supported by the occurrence of morphological adaptations in the digestive system of Prochilodus platensis, another Neotropical detritivore (Bowen, 1983). Morphological adaptations, such as long coiled intestines, are also found in Loricariidae and presumably enhance diatom digestibility (Power, 1983). The presence of such morphological adaptations would explain the large proportion of monounsaturated and polyunsaturated FAs (16:1w7, 18:1w9, 18:2w6, 20:4w6 and 22:6w3) in the muscles of P. multiradiatus when only small amounts of these FAs were recorded in the sediment of these floodplains (Mortillaro et al., 2011).

The herbivores can be divided in three subgroups. Within these subgroups, *M. aureum* has been described as feeding principally on fruits. Among fish inhabiting central Amazon Várzea, fruit feeders are common (Goulding, 1980). However, trees are not always accessible, particularly during the drought when the forest is disconnected from the lakes, and thus fruit consumption is seasonal. When fruits are inaccessible, the diet of these fish can be supplemented by plankton, detritus, macrophytes and insects (Mérona & Mérona, 2004; Oliveira et al., 2006). The ¹³C-enriched composition of M. aureum $(-27.6 \pm 0.8\%)$ and its strong proportions of $18:1\omega9$ and 18:206 separate this species from other herbivores, suggesting different feeding behaviour in the months preceding sampling (i.e. the flood season). The two other herbivores (H. microlepis and S. fasciatus) are believed to prefer algae and roots, respectively (Soares, Almeida & Junk, 1986; Silva, Ferreira & Deus, 2008). However, FA composition and stable isotope signatures suggest that they have similar diets. Previous gut content studies of S. fasciatus found its diet to be rich in plant remains (Mérona & Mérona, 2004; Oliveira et al., 2006) and roots (Soares et al., 1986). With large variability in the FA and $\delta^{13}C$ isotopic compositions of this species $(-30.3 \pm 4.1\%)$, a preference for C₃ and C₄ macrophytes cannot be discriminated. While S. fasciatus has been described as one of the rare species capable of feeding on C₄ macrophytes (Forsberg et al., 1993), large proportions of 16:1w7, 18:1w9 and 22:6w3, markers of microalgae (Napolitano, 1999; Mortillaro et al., 2011), suggest that this species unintentionally feeds on macrophyte roots while focusing on microalgae and microfauna (such as Gammaridae and Oligochaeta; Bell, Ghioni & Sargent, 1994) inhabiting the floating meadows (Engle & Melack, 1993).

In a study by Mérona & Mérona (2004), gut contents of *Hemiodopsis microlepis* (synonym for *H. microlepis*) collected in *Várzea* do Rei, a floodplain at the junction between the Amazon and Negro rivers, were composed of 50% eggs and scales. However, various studies have described this species as an omnivore (Mérona & Mérona, 2004), algivore and detritivore (Forsberg *et al.*, 1993; Silva *et al.*, 2008). In the samples collected in our study, FA composition of *H. microlepis* was indistinguishable from those of the herbivore *S. fasciatus* and the piscivore *P. nattereri*. The isotopic signature of *H. microlepis* suggests a ¹⁵N-enriched diet, comparable to an omnivore's diet. As a result, *H. microlepis* can be considered as an opportunist, like many other Amazon fish species, adapting its diet based of food availability.

Mérona & Mérona (2004) described *L. friderici* as an omnivore, with large proportions of plant remains in its gut contents. Consequently, *L. friderici* has also been reported to be a herbivore (Manetta, Benedito-Cecilio &

Martinelli, 2003), dependent upon environmental conditions (Albrecht & Caramaschi, 2003; Pouilly et al., 2003). Such herbivorous tendencies are supported by the ¹⁵N-depleted composition of L. friderici samples $(6.8 \pm 0.5\%)$, whose isotopic signature resembles those of detritivores and frugivores. However, similar FA composition was found in L. friderici and T. angulatus, an insectivore (Pouilly et al., 2003) or omnivore (Mérona & Mérona, 2004). Nevertheless, in contrast to L. friderici, the ¹⁵N-enriched composition of T. angulatus $(8.9 \pm 0.5\%)$ confirms it to be a secondary consumer. Also, as previously suggested for H. microlepis and L. friderici, the presence of terrestrial and aquatic insects, as well as fruits and flowers, in the gut contents of T. angulatus (Pouilly et al., 2003; Mérona & Mérona, 2004), indicates opportunist behaviour in these three species (M. Pouilly, pers. obs.).

Lastly, C. monoculus and P. nattereri have been reported to be piscivores (Pouilly et al., 2003; Mérona & Mérona, 2004). The ¹⁵N-enriched composition of C. monoculus and P. nattereri (9.2 \pm 0.7 and 9.9 \pm 0.4% respectively) places them as predators in the studied food web. Polyunsaturated FA composition and δ^{13} C signatures of these fish resemble those of detritivores. Strong proportions of 20:4w6 and 22:6w3 were also observed for C. monoculus by Inhamuns et al. (2009). Assuming a δ^{15} N trophic enrichment factor of 2.8% for Neotropical fish (Jepsen & Winemiller, 2007), P. multiradiatus and S. insignis, and other non-sampled species with similar isotopic signatures, may thus be considered as preferred prey for both piscivores. The pectoral spines and osseous plates of adult armoured catfish, P. multiradiatus, makes predation by C. monoculus unlikely (Nico, 2010), suggesting a stronger predation of C. monoculus on S. insignis. However, while C. monoculus is described as an exclusive piscivore (Goulding, 1980), P. nattereri may also feed on macrophytes, crabs and insects (Sazima & Machado, 1990; Pouilly et al., 2003).

Trophic structure and functioning

Our data confirm that the food web of the studied central Amazon *Várzea* is ¹³C depleted and mainly centred on a C₃ carbon source, with δ^{13} C composition ranging between -25% (-30% for C₃ macrophytes) and -40% (phytoplankton). Within this pool of primary producers, C₃ macrophytes are considered as the main carbon source for fish of Amazon floodplains (Oliveira *et al.*, 2006). For species sampled in this study, SI analyses did not detect a stronger dietary contribution of C₃ macrophytes than POM (>0.7 and 63 µm) or sediments. The

© 2015 John Wiley & Sons Ltd, Freshwater Biology, doi: 10.1111/fwb.12598

contribution of C_3 macrophytes to Amazon aquatic food webs needs to be weighted by the number of sources with similar $\delta^{13}C$ compositions (Oliveira *et al.*, 2006). Forsberg *et al.* (1993) reported that C_3 macrophytes contributed 82.4 to 97.5 % for adult fish. This discrepancy may be explained by the similar $\delta^{13}C$ compositions of terrestrial plants and periphyton pooled in the analysis by these authors. In contrast, the FA composition of species from low trophic levels, such as *S. insignis*, *P. multiradiatus* and *H. microlepis*, indicates a stronger contribution of POM with markers characteristic of microalgae.

A weak contribution of C₄ macrophytes and their associated macro- and microfauna to fish diets has been previously observed (Forsberg et al., 1993; Oliveira et al., 2006). However, C₄ macrophytes may account for up to 65% of Várzea primary production (Melack et al., 1999) and constitute the dominant proportion of the region's biomass (Junk & Piedade, 1997). Such differences between biomass production and utilisation in the food web have been tentatively attributed to low nutritional value, as well as lower digestibility due to strong fibres in C₄ macrophytes (Forsberg et al., 1993). Comparisons of the nutritional quality of C₃ and C₄ plants have produced inconsistent results (Scheirs, De Bruyn & Verhagen, 2001), which may be due to high interspecific variability. However, in the central Amazon Várzea, FA composition of the C4 P. repens displayed higher concentrations of 18:2w6 and 18:3w3 than C3 macrophyte species (Mortillaro et al., 2011). Such FAs, along with overall polyunsaturated FAs, are indicators of good organic matter quality, as they are precursors of 20:4006, 20:5w3 and 22:6w3 (Cunnane, 1996). These three FAs are essential for the metabolism and function of tissues and organs (Holman, 1998; Arts et al., 2001). In Várzea, low consumption of C₄ macrophytes may therefore be attributed to lower digestibility rather than lower nutritional value.

A spatial difference in food source utilisation is apparent for one species, S. fasciatus, in our studied food web. The δ^{13} C composition of *S. fasciatus* was enriched in Cabaliana $(-26.8 \pm 3.1\%)$ relative to Janauacá $(-33.2 \pm 2.1\%)$. This may be due to a larger contribution of C₄ macrophytes to the diet of *S. fasciatus* in Cabaliana. Consumption of C₄ carbon by species in the genus Schizodon has been previously observed (Forsberg et al., 1993; Jepsen, 1999; Manetta et al., 2003). This spatial pattern is supported by a shift in the FA composition of S. fasciatus within the floodplains, with a stronger proportion of 18:206 and 18:303 in Cabaliana. Digestive enzymes, microorganisms or digestive system modifications may allow the use of C₄ plants, despite their low digestibility for herbivores (Caswell et al., 1973). However, among the nine studied species, only S. fasciatus from Cabaliana appears to feed on C₄ macrophytes. During the sampling period, Cabaliana had greater vegetation cover (111-464 km² for flooded forest and 26-45 km² for macrophytes) than Janauacá (1-13 km² for flooded forest and 3-7 km² for macrophytes). The higher availability of flooded or aquatic vegetation in Cabaliana may explain the larger consumption of C₄ macrophytes. Therefore, the use of this carbon source may be associated with both particular environmental conditions and availability. In addition, greater total FA concentration in S. fasciatus muscles from Cabaliana suggests that these fish have a higher fat content. This finding corroborates the selection by S. fasciatus of C₄ macrophytes, as they have high nutritional value and are widely available.

Within the Amazon Várzea system, oscillation of the hydrological cycle leads to spatial and temporal variability in trophic resources (Junk, 1997), necessitating species adaptability and flexibility in diet and habitat. Species consuming seasonal resources are thought to adopt the strategy 'half a loaf is better than none'. Observations confirm the use of this strategy, as species adapt their feeding behaviours during food shortages by increasing animal food consumption (e.g. fish and invertebrates for omnivores and insectivores species; Soares et al., 1986; Wantzen et al., 2002; Mérona & Mérona, 2004). Other species, such as S. fasciatus, focus on neglected food sources, such as C₄ macrophytes, which predominate in terms of biomass. The paradox of low C₄ macrophyte consumption in the Amazon food web, however, remains largely unexplained. Our results indicate that digestibility rather than nutritional value may explain their low contribution to aquatic food webs. However, further investigation is needed to explain the fate of this predominant resource. While digestibility may be the main reason for underuse, nothing yet explains why these macrophytes are not integrated in the food web after degradation. We suggest that degradation experiments be implemented to understand the processes behind this discrepancy.

Acknowledgments

This research is a contribution to the CARBAMA project, supported by the ANR (French National Agency for Research, grant number 08-BLAN-0221) and the CNPq (National Council for Scientific and Technological Development – Brazil, Universal Program grant number 477655/2010-6). It was conducted within an international cooperation agreement between the CNPq (Brazil) and the IRD (Institute for Research and Development – France) and under the auspices of the Environmental Research Observatory Hydrology and Geochemistry of the Amazon Basin (HYBAM), supported by the INSU and the IRD. We thank Molly Moynihan for her assistance in English improvement. We are also grateful to the two anonymous reviewers for help and constructive review.

References

- Abril G., Martinez J.-M., Artigas L.F., Moreira-Turcq P., Benedetti M.F., Vidal L. *et al.* (2014) Amazon River carbon dioxide outgassing fuelled by wetlands. *Nature*, **505**, 395–398.
- Ahlgren G., Blomqvist P., Boberg M. & Gustafsson I.B. (1994) Fatty acid content of the dorsal muscle: an indicator of fat quality in freshwater fish. *Journal of Fish Biology*, 45, 131–157.
- Albrecht M.P. & Caramaschi E.P. (2003) Feeding ecology of *Leporinus friderici* (Teleostei; Anostomidae) in the upper Tocantins River, Central Brazil, before and after installation of a hydroelectric plant. *Studies on Neotropical Fauna and Environment*, **38**, 33–40.
- Alfaro A.C., Thomas F., Sergent L. & Duxbury M. (2006) Identification of trophic interactions within an estuarine food web (northern New Zealand) using fatty acid biomarkers and stable isotopes. *Estuarine Coastal and Shelf Science*, **70**, 271–286.
- Araujo-Lima C.A.R.M., Agostinho A.A. & Fabre N.N. (1995) Trophic aspects of fish communities in Brazilian rivers and reservoirs. In: *Limnology in Brazil*. (Eds J.G. Tundisi, C.E.M. Bicudo & T. Matsamura Tundisi), pp. 105–136. Brazilian Academy of Sciences/Brazilian Limnological Society, São Paulo.
- Araujo-Lima C.A.R.M., Forsberg B.R., Victoria R. & Martinelli L. (1986) Energy sources for detritivorous fishes in the Amazon. *Science*, 234, 1256–1258.
- Arts M.T., Ackman R.G. & Holub B.J. (2001) "Essential fatty acids" in aquatic ecosystems: a crucial link between diet and human health and evolution. *Canadian Journal of Fisheries and Aquatic Sciences*, **58**, 122–137.
- Bell J.G., Ghioni C. & Sargent J.R. (1994) Fatty acids compositions of 10 freshwater invertebrates which are natural food organisms of Atlantic Salmon parr (*Salmo salar*) – A comparison with commercial diets. *Aquaculture*, **128**, 301– 313.
- Benedito-Cecilio E. & Araujo-Lima C. (2002) Variation in the carbon isotope composition of Semaprochilodus insignis, a detritivorous fish associated with oligotrophic and eutrophic Amazonian rivers. *Journal of Fish Biology*, **60**, 1603–1607.
- Benjamini Y. & Hochberg Y. (1995) Controlling the false discovery rate: a practical and powerfull approach to

multiple testing. *Journal of the Royal Statistical Society Series B-Methodological*, **57**, 289–300.

- Bligh E.G. & Dyer W.J. (1959) A rapid method of total lipid extraction and purification. *Canadian Journal of Biochemistry and Physiology*, **37**, 911–917.
- Bowen S.H. (1983) Detritivory in neotropical fish communities. *Environmental Biology of Fishes*, **9**, 137–144.
- Canuel E.A., Cloern J.E., Ringelberg D.B., Guckert J.B. & Rau G.H. (1995) Molecular and isotopic tracers used to examine sources of organic matter and its incorporation into the food webs of San-Francisco bay. *Limnology and Oceanography*, **40**, 67–81.
- Caswell H., Reed F., Stephens S.N. & Werner P.A. (1973) Photosynthetic pathways and selective herbivory – hypothesis. *American Naturalist*, **107**, 465–480.
- Cunnane S.C. (1996) The Canadian Society for Nutritional Sciences 1995 Young Scientist Award Lecture. Recent studies on the synthesis, beta-oxidation, and deficiency of linoleate and alpha-linolenate: are essential fatty acids more aptly named indispensable or conditionally dispensable fatty acids? *Canadian Journal of Physiology and Pharmacology*, **74**, 629–639.
- Dalsgaard J., St John M., Kattner G., Muller-Navarra D. & Hagen W. (2003) Fatty acid trophic markers in the pelagic marine environment. *Advances in Marine Biology*, 46, 225– 340.
- Engle D.L. & Melack J.M. (1993) Consequences of riverine flooding for seston and the periphyton of floating meadows in an Amazon floodplain lake. *Limnology and Oceanography*, **38**, 1500–1520.
- Forsberg B.R., Araujolima C., Martinelli L.A., Victoria R.L. & Bonassi J.A. (1993) Autotrophic carbon sources for fish of the central Amazon. *Ecology*, 74, 643–652.
- Furch K. & Junk W. (1992) Nutrient dynamics of submersed decomposing Amazonian herbaceous plant species *Pasp-alum fasciculatum* and *Echinochloa polystachya*. *Revue d'hydrobiologie tropicale*, 25, 75–85.
- Goulding M. (1980) *The Fishes and the Forest. Exploration in Amazon Natural History.* University of California Press, Berkeley, CA.
- Goulding M., Carvalho M.L. & Ferreira E.G.F. (1988) *Rio Negro. Rich Life in Poor Water: Amazonian Diversity and Foodchain Ecology as seen Through Fish Communities.* SPB Academic Publishing, The Hague.
- Hedges J.I., Clark W.A., Quay P.D., Richey J.E., Devol A.H.
 & Santos U.D. (1986) Compositions and fluxes of particulate organic material in the Amazon river. *Limnology and Oceanography*, **31**, 717–738.
- Hess L.L., Melack J.M., Novo E.M.L.M., Barbosa C.C.F. & Gastil M. (2003) Dual-season mapping of wetland inundation and vegetation for the central Amazon basin. *Remote Sensing of Environment*, **87**, 404–428.
- Holman R.T. (1998) The slow discovery of the importance of omega 3 essential fatty acids in human health. *Journal of Nutrition*, **128**, 427S–433S.

 $\ensuremath{\mathbb{C}}$ 2015 John Wiley & Sons Ltd, Freshwater Biology, doi: 10.1111/fwb.12598

- Inhamuns A.J. & Franco M.R.B. (2001) Composition of total, neutral, and phospholipids in mapara (*Hypophthalmus* sp.) from the Brazilian Amazonian area. *Journal of Agricultural and Food Chemistry*, **49**, 4859–4863.
- Inhamuns A.J., Franco M.R.B. & Batista W.S. (2009) Seasonal variations in total fatty acid composition of muscles and eye sockets of tucunare (*Cichla* sp.) from the Brazilian Amazon area. *Food Chemistry*, **117**, 272–275.
- Izquierdo M.S. (1996) Essential fatty acid requirements of cultured marine fish larvae. *Aquaculture Nutrition*, 2, 183– 191.
- Jepsen D.B. (1999) Analysis of Trophic Pathways in Freshwater Ecosystems Using Stable Isotope Signatures. Texas A & M University, College Station, Texas, USA.
- Jepsen D.B. & Winemiller K.O. (2007) Basin geochemistry and isotopic ratios of fishes and basal production sources in four neotropical rivers. *Ecology of Freshwater Fish*, **16**, 267–281.
- Junk W.J. (1985) The Amazon floodplain: a sink or source for organic carbon. Mitteilungen aus dem Geologische-Palaontologischen Institut der Universitat Hamburg. SCOPE/ UNEP Sonderband, 58, 267–283.
- Junk W.J. (1997) The Central Amazon Floodplain: Ecology of a Pulsing System. Springer, Berlin, Heidelberg, New York.
- Junk W.J. & Howard-Williams C. (1984) Ecology of aquatic macrophytes in Amazonia. In: *The Amazon, Limnology and Landscape Ecology of a Mighty Tropical River and its Basin.* (Ed. H. Sioli), pp. 269–293. Junk, Dordrecht.
- Junk W.J. & Piedade M.F.T. (1997) Plant life in the floodplain with special reference to herbaceous plants. In: *The central Amazon floodplain: Ecology of a pulsing system*. (Ed. W.J. Junk), pp. 147–185. Springer, Berlin, Heidelberg, New York.
- Junk W.J. & Piedade M.T.F. (1993) Biomass and primary production of herbaceous plant communities in the Amazon floodplain. *Hydrobiologia*, **263**, 155–162.
- Koussoroplis A.M., Bec A., Perga M.E., Koutrakis E., Bourdier G. & Desvilettes C. (2011) Fatty acid transfer in the food web of a coastal Mediterranean lagoon: evidence for high arachidonic acid retention in fish. *Estuarine Coastal and Shelf Science*, **91**, 450–461.
- Koven W.M., Kissil G.W. & Tandler A. (1989) Lipid and η-3 requirement of *Sparus aurata* larvae during starvation and feeding. *Aquaculture*, **79**, 185–191.
- Kwetegyeka J., Masa J., Kiremire B.T., Mpango G.B. & Grahl-Nielsen O. (2011) Fatty acids of polar lipids in heart tissue are good taxonomic markers for tropical African freshwater fish. *African Journal of Aquatic Science*, **36**, 115–127.
- Lowe-McConnell R.H. (1987) *Ecological Studies in Tropical Fish Communities*. Cambridge University Press, Cambridge.
- Manetta G.I., Benedito-Cecilio E. & Martinelli M. (2003) Carbon sources and trophic position of the main species

© 2015 John Wiley & Sons Ltd, Freshwater Biology, doi: 10.1111/fwb.12598

of fishes of Baía River, Paraná River floodplain, Brazil. *Brazilian Journal of Biology*, **63**, 283–290.

- Marshall B.G., Forsberg B.R. & Thome-Souza M.J.F. (2008) Autotrophic energy sources for *Paracheirodon axelrodi* (Osteichthyes, Characidae) in the middle Negro River, Central Amazon, Brazil. *Hydrobiologia*, **596**, 95–103.
- Melack J.M., Forsberg B.R., Victoria R.L. & Richey J.E. (1999) Biogeochemistry of Amazon floodplain lakes and associated wetlands. In: *The Biogeochemistry of the Amazon Basin and its Role in a Changing World* (Eds M. Mcclain, R. Victoria & J. Richey), pp. 1–50. Oxford University Press, Oxford, United Kingdom.
- Merigoux S. & Ponton D. (1998) Body shape, diet and ontogenetic diet shifts in young fish of the Sinnamary River, French Guiana, South America. *Journal of Fish Biology*, **52**, 556–569.
- Mérona B. & Mérona J.R. (2004) Food resource partitioning in a fish community of the central Amazon floodplain. *Neotropical Ichthyology*, **2**, 75–84.
- Mortillaro J.M., Abril G., Moreira-Turc P., Sobrinho R., Perez M. & Meziane T. (2011) Fatty acid and stable isotope (δ^{13} C, δ^{15} N) signatures of particulate organic matter in the Lower Amazon River: seasonal contrasts and connectivity between floodplain lakes and the mainstem. *Organic Geochemistry*, **42**, 1159–1168.
- Napolitano G.E. (1999) Fatty acids as trophic and chemical markers in freshwater ecosystems. In: *Lipids in Freshwater Ecosystems* (Eds M.T. Arts & B.C. Wainman), pp. 21–44. Springer-Verlag, New York, USA.
- Nico L.G. (2010) Nocturnal and diurnal activity of armored suckermouth catfish (Loricariidae: *Pterygoplichthys*) associated with wintering Florida manatees (*Trichechus manatus latirostris*). *Neotropical Ichthyology*, **8**, 893–898.
- Oksanen J., Blanchet F.G., Kindt R., Legendre P., O'Hara R.B., Simpson G.L. *et al.* (2010) vegan: Community Ecology Package. R package version 1.17-12.
- Oliveira A.C.B., Soares M.G.M., Martinelli L.A. & Moreira M.Z. (2006) Carbon sources of fish in an Amazonian floodplain lake. *Aquatic Sciences*, **68**, 229–238.
- Parrish C.C., Abrajano T.A., Budge S.M., Helleur R.J., Hudson E.D., Pulchan K. *et al.* (2000) Lipid and phenolic biomarkers in marine ecosystems: analysis and applications. In: *The Handbook of Environmental Chemistry* (Ed. P. Wangersky), pp. 193–223, Vol. D, Marine Chemistry. Springer-Verlag, Berlin, Heidelberg.
- Peterson B.J. & Fry B. (1987) Stable isotopes in ecosystem studies. *Annual Review of Ecology and Systematics*, **18**, 293–320.
- Piedade M.T.F., Junk W.J. & Long S.P. (1991) The productivity of the C₄ grass *Echinochloa polystachya* on the Amazon floodplain. *Ecology*, **72**, 1456–1463.
- Pouilly M., Lino F., Bretenoux J.G. & Rosales C. (2003) Dietary-morphological relationships in a fish assemblage of the Bolivian Amazonian floodplain. *Journal of Fish Biology*, 62, 1137–1158.

12 J. M. Mortillaro et al.

- Power M.E. (1983) Grazing responses of tropical freshwater fishes to different scales of variation in their food. *Environmental Biology of Fishes*, **9**, 103–115.
- Quay P.D., Wilbur D.O., Richey J.E., Hedges J.I., Devol A.H. & Victoria R. (1992) Carbon cycling in the Amazon River – Implications from the ¹³C compositions of particles and solutes. *Limnology and Oceanography*, **37**, 857–871.
- Rainuzzo J.R., Reitan K.I. & Olsen Y. (1997) The significance of lipids at early stages of marine fish: a review. *Aquaculture*, **155**, 103–115.
- Sazima I. & Machado F.A. (1990) Underwater observations of Piranhas in western Brazil. *Environmental Biology of Fishes*, **28**, 17–31.
- Scheirs J., De Bruyn L. & Verhagen R. (2001) A test of the C_3 - C_4 hypothesis with two grass miners. *Ecology*, **82**, 410–421.
- Silva C.C., Ferreira E.J.G. & Deus C.P. (2008) Diet of five species of Hemiodontidae (Teleostei, Characiformes) in the area of influence of the Balbina reservoir, Uatuma River, State of Amazonas, Brazil. *Iheringia Serie Zoologia*, 98, 464–468.
- Sioli H. (1984) The Amazon, Limnology and Landscape Ecology of a Mighty Tropical River and its Basin, Junk, Dordrecht.
- Soares M.G.M., Almeida R.G. & Junk W.J. (1986) The trophic status of the fish fauna in lago Camaleao, a macrophyte dominated floodplain lake in the middle Amazon. *Amazoniana-Limnologia Et Oecologia Regionalis Systemae Fluminis Amazonas*, 9, 511–526.

- Townsend-Small A., Mcclain M.E. & Brandes J.A. (2005) Contributions of carbon and nitrogen from the Andes Mountains to the Amazon River: evidence from an elevational gradient of soils, plants, and river material. *Limnology and Oceanography*, **50**, 672–685.
- Wantzen K.M., Machado F.D., Voss M., Boriss H. & Junk W.J. (2002) Seasonal isotopic shifts in fish of the Pantanal wetland, Brazil. *Aquatic Sciences*, 64, 239–251.
- Wetzel R.G. (2001) *Limnology: Lake and River Ecosystems*. Academic press, San Diego, California, USA.
- Yamamoto K.C., Soares M.G.M. & Freitas C.E.C. (2004) Alimentação de *Triportheus angulatus* (Spix & Agassiz, 1829) no Camaleão, Manaus, AM, Brasil. *Acta Amazonica*, 34, 653–659.
- Zuur A.F., Ieno E.N. & Smith G.M. (2007) *Analysing Ecologi*cal Data. Springer, Heidelberg, Germany.

Supporting Information

Additional Supporting Information may be found in the online version of this article:

Table S1. FAs composition of leaves and roots of 4 macrophytes species.

Table S2. FAs composition of POM (>0.7 and 63 $\mu m)$ and sediment samples.

(Manuscript accepted 8 April 2015)