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Effect of *in vitro* exposure to zinc on immunological parameters of haemocytes from the marine gastropod *Haliotis tuberculata*

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ABSTRACT

Environmental pollutants such as heavy metals exert immunotoxic effects on aquatic organisms. The immune defence of molluscs is comprised of cell-mediated and humoral mechanisms, in which haemocytes play a key role. In this study, a model based on primary cultured haemocytes from the gastropod mollusc *Haliotis tuberculata* was established to investigate the effects of zinc *in vitro*. Cells were exposed for 24 h to ZnCl₂ concentrations of 0, 10, 100 or 1000 μ M. The effects of zinc on haemocyte parameters were investigated using morphological, spectrophotometric and flow cytometry analysis. Immunotoxicity was reflected by a significant decrease in the number of viable haemocytes (LC₅₀(24 h) = 314 μ M). Moreover, the cell area was dramatically reduced, and the percentage of rounded cells increased with increasing zinc concentrations. Exposure to 1000 μ M zinc induced a significant reduction in acid phosphatase activity, phagocytic activity and reactive oxygen species production in haemocytes. However, several haemocyte parameters increased significantly after 24 h of zinc exposure. In response to a 1000 μ M exposure, the phenoloxidase level was 26-fold higher than that of the control, and nonspecific esterase activity was increased by 69% above that of the control. These results suggest a relationship between zinc exposure and alterations in the functional responses of haemocytes from *H. tuberculata*.

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1. Introduction

Heavy metals are continually discharged into the marine environment by human activities. Anthropogenic sources of metals include urban runoff, sewage, traffic emissions, coal and oil combustion, and industrial production [1]. Among these metals, zinc is one of the most widely spread environmental toxicants [2].

Zinc is an essential trace element for all living organisms. This metal is a constituent of a large number of zinc-dependent enzymes that are important for cellular function and zinc plays a crucial role in stabilising membranes [2–6].

Zinc is also known to play a central role in the immune system. In vertebrates, this metal is crucial for the normal development and functioning of cells that mediate immunity, such as neutrophils and natural killer cells [7]. For example, zinc deficiency affects the development of acquired immunity by preventing both the outgrowth and various functions of T lymphocytes, such as activation, Th1 cytokine production, and B lymphocyte assistance. Macrophages, which are pivotal cells in many immunological functions (such as intracellular killing, cytokine production and phagocytosis), and their activation can be deregulated by zinc deficiency [7]. However, zinc can also be potentially toxic to organisms if it occurs at high concentrations [8]. When present in excess, Zinc can have deleterious effects on the immune system of vertebrates [9].

In molluscs, the cellular defence system is represented by haemocytes, which are responsible for phagocytosis, pathogen hydrolysis, production of reactive oxygen species and the phenoloxidase cascade [10-13]. Numerous studies have been conducted on bivalve haemocytes, and these studies demonstrate that haemocytes functions (cell viability, cytoskeletal organisation, phagocytic activity) can be affected by several contaminants, including

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heavy metals [11,14–18]. In gastropods, and in particular in abalone, several studies have reported that haemocytes can be affected by environmental factors such as abiotic stresses [19–21], infections [22] and organic compounds [23], but little is known about the effect of heavy metals.

To assess the effect of zinc on abalone immune response, an *in vitro* approach was chosen. During the past decade, invertebrate cell culture had been developed to study physiological processes [24–27] and pollutant effects [28]. In ecotoxicology, this approach is an alternative to animal testing because of the reduced use of experimental animals, low cost and rapid performance [29]. Moreover, an *in vitro* approach allows the use of specific endpoints to determine the targets of toxic effects with great precision and reproducibility. The major disadvantage of *in vitro* analysis systems is the difficulty of extrapolating the results to the *in vivo* condition.

Although numerous studies have already been conducted on the toxic effects of zinc on marine animals [30,31] and, more specifically, on molluscs [16,18], few investigations have reported the effects of heavy metals on marine gastropods at the cellular level. The aim of the present study was to investigate the *in vitro* effects of zinc on several haemocyte parameters in the European abalone, *Haliotis tuberculata*. Cell viability and morphology, phagocytic activities, and production of reactive oxygen species were analysed.

2. Materials and methods

2.1. Animals

Living adult abalone, *H. tuberculata*, 9–11 cm in shell length, originating from the northern Cotentin peninsula (France) were purchased. They were then maintained in natural and continuously aerated seawater at 17 °C at the Centre de Recherche en Environnement Côtier (C.R.E.C., Luc-sur-Mer, Lower-Normandy, France). The studied abalones were regularly fed with a mixture of *Laminaria sp.* and *Palmaria palmata*.

2.2. Primary cell cultures

Haemocytes were cultured as previously described [24,26,32]. Briefly, after an incision in the foot, haemolymph was collected (10-15 ml per animal) using a 20-ml syringe fitted with a 25-gauge hypodermic needle. Haemolymph was transferred to a sterile tube and immediately diluted 1:4 in cooled, sterile, anti-coagulant modified Alsever's solution (115 mM glucose, 27 mM sodium citrate, 11.5 mM EDTA, 382 mM NaCl) [33]. Haemocytes were rapidly plated at 1.5.10⁶ cells in 6-wells plates (MTT assay or enzymatic analysis) or at 0.5.10⁶ cells in 12-wells plates (flow cytometry analysis) into which three volumes of sterile artificial water were added. Cultures were maintained at 17 °C in an incubator (CO₂ free). After 90 min of incubation, cells were covered with Hank's sterile 199 medium modified by the addition of 250 mM NaCl, 10 mM KCl, 25 mM MgSO₄, 2.5 mM CaCl₂, 10 mM Hepes, 2 mM L-glutamine, 100 µg/ml streptomycin and 60 µg/ml penicillin G at a final pH of 7.4. The cells were then kept at 17 °C for 24 h before beginning the experiments. Then, cells were exposed for 24 h to a range of ZnCl₂ concentrations: 0, 10, 100, or 1000 µM. Zinc concentrations ranging from 7-46 µM (0.47-3000 µg Zn/L) have been reported in coastal waters worldwide that are inhabited by Haliotis spp. [34].

2.3. MTT assay

Cell viability was evaluated by an MTT (3-[4,5-dimethylthiazol-2-yl]-2,5-diphenyltetrazolium bromide) reduction assay [35] adapted to molluscan cell cultures [24,36] and $LC_{50}(24 h)$ was

used as an index of cytotoxicity. The MTT assay is a colorimetric metabolic assay based on mitochondrial dye conversion to assess viability. Briefly, 10% (v/v) of the MTT stock solution (5 mg MTT/ml of PBS) was added to the culture dishes. After 24 h of incubation, an equal volume of isopropanol containing 0.04 N HCl was added to each culture to dissolve the converted formazan. Absorbance was then measured at a wavelength of 570 nm with a 630 nm reference.

2.4. Analysis of cellular morphology

To test the potential effects of zinc on cell morphology, haemocytes were cultured as described above on coverslips. Haemocyte monolayers were fixed in a solution of methanol for 10 min at room temperature. The coverslips were placed for 1 min in hematoxylin to stain nuclei, then briefly rinsed in distilled water and incubated for 1 min in light green dye to stain the cytoplasm. Haemocytes were then incubated for 10 min in acetone, and finally, the coverslips were mounted in Roti-Histol[®] (Roth, Karsruhe, Germany). Observations were carried out with a Nikon Eclipse 80i light microscope. The cell and nuclear areas were measured using a computer-assisted microscopic image analysis system, the NISelements D 2.30 software (Nikon).

2.5. Flow cytometry analysis of abalone haemocytes

Haemocyte analysis was performed using an EPICS XL 4 (Beckman Coulter), and 20 000 events were counted for each sample. Results are expressed as cell cytograms indicating the size (FSC value), the complexity (SSC value) and the level of fluorescence using the FL1 channel.

2.5.1. Phagocytosis capacity

Phagocytosis activity was measured by quantifying the ingestion of fluorescent beads (carboxylate-modified FluoroSpheres[®], yellow-green fluorescence, 1 μ m diameter, Molecular Probes). In each culture well, 7 μ L of bead solution were added, and cells were incubated for 60 min at 17 °C in the dark.

2.5.2. Reactive oxygen species (ROS) production

ROS production was evaluated using the 2',7'-dichlorofluorescein diacetate (DCFH-DA, sigma) method [37] as adapted to mollusc cells by Lambert et al. [38]. Haemocytes were incubated for 20 min at 17 °C in the dark with DCFH-DA to a final concentration of 100 μ M. Then, 6 μ L of phorbol 12-myristate 13-acetate (PMA, 1 mg/ml, Sigma) were added to each well, and cells were incubated for 60 min at 17 °C in the dark.

2.5.3. non-specific esterase activity

Esterase activity was measured using the non-specific liposoluble substrate fluorescein diacetate (FDA, Molecular Probes). The FDA probe was added to each well at a final concentration of 5 μ M, and cells were incubated for 60 min at 17 °C in the dark.

2.5.4. Cell fixation

After the incubation, the wells were gently scraped, and haemocyte samples were centrifuged ($500 \times g$, 10 min, 4 °C). Then, the supernatants were removed, and cells were fixed with paraformaldehyde (2% in distilled water). Samples were stored at 4 °C until analysis.

The results of the phagocytosis experiments are expressed as percentages of haemocytes containing three beads or more, according to Delaporte et al. [39], Hegaret et al. [40] and Labreuche et al. [41]. For the ROS production experiments, lysosome-positive cell assays and non-specific esterase activity studies, results are expressed as the percentage of cells containing fluorescence.

2.6. Enzymatic activities

Acid phosphatase (AcP) activity was evaluated as previously described by González et al. [42]. Briefly, the culture medium was removed, and the cells were scraped and lysed in 50 mM Tris buffer (pH 7.4). An equal volume of 1.5 M Tris buffer (pH 5) supplemented with 5 mM *p*-NPP (*para*-nitrophenyl phosphate, Sigma) as a substrate, 1 mM ZnCl₂, and 1 mM CaCl₂ were added to each sample. The samples were then incubated at 37 °C for 30 min. The reaction was stopped with 0.3 N NaOH, and the absorbance was measured at 410 nm. The AcP activity was estimated with a standard curve of *p*-nitrophenol.

Detection of phenoloxidase activity in the haemocyte culture medium was carried out by measurement of 3,4-dihydroxy-L-phenylalanine (L-DOPA) transformation in dopachromes [43]. Transformation was monitored by spectrophotometry at 492 nm. Briefly, 250 μ l of Tris (10 mM)/NaCl (150 mM) buffer at pH 8, 250 μ l of cacodylate buffer (10 mM) and 250 μ l of L-DOPA (3 mg/ml) were added to a 250 μ l sample. Absorbance was then measured after 1 h of incubation at 25 °C. Phenoloxidase activity was expressed as an arbitrary unit.

The amount of protein in each cellular extract and in the medium of each primary culture was determined by the Bradford method [44] using bovine serum albumin as a standard.

2.7. Data analysis

Results are expressed as means \pm S.D. Each experiment was repeated at least three times, and the means were calculated from triplicates for each experiment. The significance of the differences between mean values was estimated using student's *t*-tests

3. Results

3.1. Effects of zinc on haemocyte viability

A dose–response cytotoxicity curve, as quantitated by an MTT assay after 24 h of exposure of haemocytes to zinc, is shown Fig. 1. No significant change was observed at 10 μ M Zn. At zinc concentrations of 100 μ M and higher, haemocyte viability decreased, and maximum toxicity was observed when cells were exposed to 1000 μ M zinc. The LC₅₀(24 h) value, was determined graphically from the semilogarithmic plots representing the zinc concentration required to reduce cell viability to 50%; the LC₅₀(24 h) for Zinc was approximately 314 μ M for haemocytes in culture.



Fig. 1. Dose-response effect of zinc on cell viability as determined by the MTT reduction assay. Haemocytes were plated at a density of 1.5.10⁶ cells per well in 6-well culture plates and cultured at 17 °C in modified Hank's-199 medium. 24 h after the beginning of cultures, the medium was renewed. Cells were exposed for 24 h to ZnCl₂ and MTT test was performed. Each data point represents the mean percentage \pm standard deviation of triplicate cultures. Significant difference from control cells at *P* < 0.001 (***).

3.2. Effects of zinc on haemocyte morphology

The morphological effect of zinc on haemocytes after 24 h of exposure is illustrated in Fig. 2. Haemocytes were noticeably affected by the presence of zinc. In the absence of zinc (control), cells were interconnected and elongated with large pseudopods. In the presence of zinc, haemocytes became isolated and more rounded. This change was quantified by the measurement of haemocyte area, nucleocytoplasmic ratio and a shape factor (Fig. 3). In the control preparation of Haliotis haemocytes, the mean cellular area measured $210.4 \pm 5.5 \,\mu\text{m}^2$ (range $84.3 - 384.6 \,\mu\text{m}^2$). The means of haemocyte area in zinc-treated cells decreased in a dosedependent manner, resulting in cell areas that were decreased by 36% (133.4 \pm 4.6 μm^2 vs. 210.4 \pm 5.5 $\mu m^2)$ and by 52% $(99.9\pm3.4\,\mu m^2$ vs. $210.4\pm5.5\,\mu m^2)$ in the presence of 100 μM and 1000 µM zinc, respectively. Concurrently, the nucleocytoplasmic ratio significantly increased by 17% (0.23 \pm 0.01 vs. 0.20 \pm 0.01) and 36% (0.27 \pm 0.01 vs. 0.20 \pm 0.01) when haemocytes were exposed to 100 µM and 1000 µM zinc, respectively.

3.3. Zinc effects on haemocyte cellular activity

The influence of zinc on non-specific esterase and phagocytic activities as well as ROS production was assessed by flow cytometry after 24 h of exposure (Fig. 4). The percentage of haemocytes presenting phagocytic and a non-specific esterase activities as well as ROS production were $32.6 \pm 5\%$, $34.9 \pm 2\%$ and $31.5 \pm 4\%$, respectively, for each control. At low concentration (10 μ M), zinc stimulated esterase and phagocytic activities slightly, but not significantly. Compared to the control, the highest zinc concentration (1000 μ M) decreased phagocytic activity and ROS production by 22% (78.82 \pm 3.08% in treatment vs. 100% for the control) and 35% (65.65 \pm 12.01% in treatment vs. 100% for the control), respectively. For non-specific esterase activity, the percentage of fluorescent cells increased by 69% (169.60 \pm 5.07% in treatment vs. 100% for the control) after exposure to 1000 μ M zinc.

3.4. Zinc effects on haemocyte enzymatic activity

After 24 h of zinc exposure, phenoloxidase activity was stimulated in a dose-dependent manner compared to the control. A 6-fold and a 26-fold induction of phenoloxidase activity were observed with treatments of 100 μ M and 1000 μ M zinc, respectively (Fig. 5A). For AcP activity, no significant effects were recorded after an incubation with 10 μ M or 100 μ M zinc. When cells were exposed to 1000 μ M zinc, a decrease in AcP activity of 77% was observed in comparison with the control (Fig. 5B).

4. Discussion

The results obtained in this study demonstrate the toxic effects of zinc on a primary culture of haemocytes from the gastropod *H. tuberculata*. After 24 h of exposure, the observed viability was not more than of 68% and 33% compared to the control for treatments of 100 μ M and 1000 μ M ZnCl₂, respectively. Decreases in the viability of various cell types, including haemocytes from *Mya arenaria*, coelomocytes from *Lumbricus terrestris*, interrenal cells from the trout *Oncorhynchus mykiss* and cells of the human lineage ARPE-19, have been reported after 18 h or 24 h of cell exposure to zinc at concentrations ranging from 10⁻⁵ M to 10⁻³ M [45,46]. Zinc is not the only metallic compound able to induce mortality in haemocytes in primary culture. For example, AgNO₃, CdCl₂, CH₃HgCl and HgCl₂ also cause a decrease in the viability of haemocytes from several bivalves [16,45]. In our study, the LC₅₀ value (representing the zinc concentration at which 50% of cells died after



Fig. 2. Light microscopic pictures showing morphology of *Haliotis tuberculata* haemocytes. Cells were seeded on coverslips at 0.5.10⁶ cells per well and grown at 17 °C in modified Hank's-199 medium in absence (A) or presence to a range of zinc chloride concentrations: 10 μM (B), 100 μM (C), 1000 μM (D). Cells were exposed for 24 h to ZnCl₂

24 h of exposure) was calculated at 314 μ M. Brousseau et al. [45] reported a decrease of more than 50% in the viability of haemocytes from *M. arenaria* exposed *in vitro* for 18 h to ZnCl₂ at concentrations between 100 and 1000 μ M. Thus, our results appear to be in agreement with previous reports.

In addition to haemocyte viability, the extension of pseudopods by haemocytes and their capacity to adhere play key roles in cellular migration and the processes of immune defence of molluscs. The morphological results obtained in this study highlight significant physical changes in the cells exposed to zinc. The morphological changes observed with increasing concentrations of zinc were characterised by both a decrease in the spreading of haemocytes and a decrease in the number of cytoplasmic extensions. A concomitant increase in the number of rounded cells and a decrease in the cell area values (by \sim 52%, 1000 μ M zinc vs. control) were observed. These modifications were accompanied by an increase of the nucleocytoplasmic ratio. These results are in agreement with data from the literature. For example, Olabarrieta et al. [15] and Gómez-Mandikute and Cajaraville [47] showed that haemocytes from mussels (*Mytilus galloprovincialis*) in primary culture exposed to cadmium or to copper became rounded and did not present additional cytoplasmic extensions when the concentration of the metals increased. Such morphological modifications of haemocytes in response to metallic treatments are commonly associated with disturbances of the cytoskeleton, such as an intracellular disorganisation of actin [14,15].

The analysis by flow cytometry allowed us to study complementary immune parameters, such as phagocytosis, lysosomepositive cells, production of ROS and non-specific esterase activity. Phagocytosis is generally considered as the first line of defence in molluscs. We showed that phagocytic activity decreased when the haemocytes were cultured in the presence of 100 μ M Zn. These results are in agreement with previously reported studies that showed, in primary culture, an inhibition of phagocytic activity in the haemocytes of several species of bivalves, including *M. arenaria*, *Crassostrea virginica*, *Mytilus edulis*, exposed to different metallic



Fig. 3. Effect of zinc on haemocyte morphological parameters after 24 h exposure. Variation in cellular area (A), nucleocytoplasmic ratio (B) and in cell shape factor (C). Each data point represents the mean percentage \pm standard deviation of triplicate cultures. Significant difference from control cells at P < 0.001 (***).



Fig. 4. Effect of zinc on immune parameters after 24 h incubation. Phagocytic activity (A), lysosome-positive cells (B), ROS production (C), and non-specific esterase activity (D). Each data point represents the mean percentage \pm standard deviation of triplicate cultures. Significant differences from control cells at P < 0.05 (*), P < 0.001 (***).

components, such as Ag, Al, Cd, Cu, Hg [18,51] and especially ZnCl₂ [16,45,48].

A decrease in phagocytic activity can be related to a disorganisation of the cytoskeleton, with an inhibition of the mobility



Fig. 5. Effect of zinc on enzymatic activity after 24 h incubation. Phenoloxidase activity (A), Acid phosphates activity (B). Each data point represents the mean percentage \pm standard deviation of triplicate cultures. Significant difference from control cells at *P* < 0.01 (**), *P* < 0.001 (***).

of haemocytes as a consequence. Indeed, *in vitro* studies applied to bivalve cells showed that morphological and functional modifications of haemocytes could be responsible for the observed decrease in the phagocytic activity [14,49–51]. In the present study, the increase of the proportion of rounded cells observed during treatment with zinc suggests a direct or indirect effect of zinc in the disorganisation of the cytoskeleton that is characterised by a decrease of the phagocytic activity of haemocytes.

Flow cytometry allowed us to follow the production of reactive oxygen species (ROS). As was observed for phagocytosis, zinc induced an inhibition of ROS production, which, at a concentration of 1000 µM zinc, was reduced by 34% of that in the control. Gómez-Mendikute and Cajaraville [47] also noticed a decrease in ROS production by haemocytes of M. galloprovincialis exposed in vitro to Cu for 24 h. Our study was in agreement with these results. However, these same authors indicated no modification of the ROS production by these cells during exposures to Cd. On the other hand, several authors observed an increase of the ROS production by different cell types exposed to Zn in culture. As an example, 100 µM Zn^{2+} cation stimulated ROS production in cells of the human lineage ARPE-19 treated for 24 h [46]. Similarly, the production of O²⁻ was stimulated in haemocytes from M. galloprovincialis exposed to 50 µM ZnCl₂ [51]. The difference between these latter results and ours could be explained by the different exposure times; the Zn exposure time was only 30 min for the experiments carried out by Kaloyianni et al. [51], whereas it was 24 h in our study.

Our *in vitro* experiments exposing cultured cells to Zn demonstrate a correlation between phagocytic activity, morphological changes and the ROS production. The decrease in ROS production could be explained by a disorganisation of the cytoskeleton. As suggested by Thiagarajan et al. [52], the disturbance of the cytoskeleton could be correlated to the NADPH-oxidase complex contained in the plasma membrane. This complex is responsible for the formation of O^{2-} in haemocytes and could be inhibited in response to a treatment by zinc.

The presence of zinc (1000 μ M) in the cellular culture medium significantly stimulated non-specific esterase activity in the haemocvtes. The activity of these enzymes, which are involved in the intracellular degradation of foreign bodies, was studied by flow cytometry. Our results are in agreement with those obtained by Bado-Nilles et al. [53] who observed a stimulation of non-specific esterase activity in haemocytes from Crassostrea gigas that were exposed in vitro to naphthalene (HAP). However, Gagnaire et al. [17,54] showed that the non-specific esterase activity was not modified by mercury, but it was inhibited by some pesticides (paraoxon and chlorohalonil) and HAP (benzo[a]pyrène, phenanthrene, anthracene and fluoranthene). Mazón et al. [55] and Zvereva et al. [56] reported an in vitro inhibition of the non-specific esterase activity in cells (exposed to Ni, Cd and Cu) from Littorina littorea and Chrysomela lapponica. The results published in the literature on non-specific esterase activity report different effects according to the studied metallic species. Zinc is an essential metal necessary for the function of numerous enzymes including some esterases, such as phosphodiesterases [57]. Therefore, it is possible that the activity of these enzymes may be stimulated by the introduction of metal into the culture medium.

Our enzymatic analyses focussed on the activity of phenoloxidase (PO) and acid phosphatase (AcP). Phenoloxydase is a copperdependent enzyme directly implicated in mechanisms of immune defence. PO has been detected in the haemolymph of several invertebrates [58-60]. To our knowledge, the pro-phenoloxydase system has not vet been characterised in abalones. However, recent studies relative to the effects of environmental parameters and pathogens in the immune response of Haliotis diversicolor supersexta and H. tuberculata measured PO activity [19-21]. These studies indicated that abiotic factors such as temperature, salinity, and levels of ammonium or nitrite present in the water, as well as biotic factors (such as spawning) induced a decrease of PO activity. Our study showed that PO activity increased in a dose-dependent manner when haemocytes were exposed to zinc. In the literature, few studies have reported an increase in the PO activity of haemocytes in culture in the presence of anthropogenic contaminants. Nevertheless, Bado-Nilles et al. [53] showed an increase in PO activity in vitro in haemocytes from C. gigas exposed for 24 h to various concentrations of benzo[b]fluoranthene. In the ascidian Styela plicata, the in vitro exposure of haemocytes to 1, 10, 100 or 1000 µg/L of ZnSO₄ for 72 h did not affect the PO activity [61], whereas the Zn^{2+} cation (5 mM) inhibited the activity of this enzyme in the crab Charybdis japonica [59]. Mercury also inhibited the PO activity of haemocytes from C. gigas in culture after 21 h of exposure [17]. Therefore, the response of PO activity to contaminants appears to vary according to the animal species and the type and form of contaminants.

Contrary to the PO activity, the activity of acid phosphatase was inhibited by zinc. This inhibition was by more than 70% of the control activity for the highest concentration. These results are in agreement with those obtained by Mazorra et al. [62] who also demonstrated a decrease in the activity of this enzyme from Scrobicularia plana exposed for 48 h to different heavy metals, including Cd (0.05–1 mM), Cu (0.1–1 mM), Hg (0.05–1 mM) and particularly Zn (0.05–1 mM). Some metals are capable of stimulating the activity of phosphatases. Thus, the AcP activity is up-regulated in haemocytes from M. galloprovincialis after in vitro exposure to numerous concentrations of Cd [15]. A similar result was obtained in gills and digestive glands from Pinctada fucata exposed to copper [63]. However, the response of phosphatase to metallic contamination can differ according to the organ in which it is measured. For example, Rajalakshmi and Mohandas [64] observed a stimulation of AcP activity in gills and an inhibition of this activity in the digestive gland from *Lamellidens corrianus* exposed to copper.

Acid phosphatase is a lysosomal enzyme involved in the degradation of pathogenic or foreign elements. Therefore, the decrease of its activity in the presence of metals, and particularly of zinc, could be due either to a specific inhibition and/or to a more general change of the lysosomal system. Recent studies have been focused on the analysis of AcP activity in the haemolymph of Haliotis diversicolor supertexta infected by pathogenic bacteria [65]. In this study. AcP activity was increased in the haemolymph of abalone infected by Vibrio parahaemolyticus. Such a result shows that this enzyme might be secreted into the extracellular fluid by haemocytes in response to bacterial infection. In our work, AcP activity was analysed at the cellular level. Therefore, it is possible that the modulations of haemocytes exposed to a strong zinc concentration that we observed are caused by the same processes as in the case of infection by a pathogenic organism. Thus, the decrease measured during our experiments would reflect a more important secretion of this enzyme by haemocytes and a decrease of the stock of this protein inside haemocytes.

5. Conclusions

In summary, our in vitro experiments demonstrated that contamination by zinc could induce important physiological effects in haemocytes. Exposure to 100 µM Zn decreased cell viability and induced morphological and physiological modifications in haemocytes. Furthermore, this contamination decreased ROS production and increased PO activity. For the highest concentration we tested (1000 µM Zn), exposure induced an inhibition of phagocytic capacity, a decrease in acid phosphatase activity, and an increase in the non-specific esterase activity in addition to the effects already observed at 100 µM. Except for cell viability and morphology, the most sensitive haemocyte parameters to zinc are PO activity and ROS production. It was not surprising to observe an association between PO and ROS production. Indeed, it has been established that, to fight against infections, the activation of PO catalyses the production of ROS during the processes of repair, wound healing or encapsulation [66].

Increased mortality of haemocytes is associated with a change of cell shape, which becomes rounded, a massive secretion of PO, a decrease of AcP activity and an inhibition of phagocytic capacity and ROS production. Zinc is a trace element essential to cell functions, but it was shown in this study that it could become toxic if present in high levels (100 and 1000 μ M). Thus, zinc induced a general disturbance of the immune capacities of haemocytes. In conclusion, it can be speculated that environmental contamination by zinc can render abalone more vulnerable to infections or environmental stresses.

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References

- Zhou Q, Zhang J, Fu J, Shi, Jiang G. Biomonitoring: an appealing tool for assessment of metal pollution in the aquatic ecosystem. Anal Chim Acta 2008;606:135–50.
- [2] Eisler R. Zinc hazards to fish, wildlife and invertebrates: a synoptic review. Biological report 10, US Department of interior. Washington, DC: Fish and Wildlife Service; 1993.
- [3] Sandstrom B. Bioavailability of zinc. Eur J Clin Nutr 1997;51:17-9.

- [4] Chvapil M. New aspects in the biological role of zinc: a stabilizer of macromolecules and biological membranes. Life Sci 1973;13:1041-9.
- [5] Stohs SJ, Bagghi D. Oxidative mechanisms in the toxicity of metal ions. Free Radic Biol Med 1995;18:321–36.
- [6] Frassinetti S, Bronzetti G, Caltavuturo L, Cini M, Della Croce C. The role of zinc in life: a review. J Environ Pathol Toxicol Oncol 2006;25:597–610.
- [7] Ibs K-H, Rink L. Zinc-altered immune function. J Nutr 2003;133:1452S-6S.
- [8] Harris ED. Biological monitoring of iron, zinc and copper. In: Dillon HK, Ho MH, editors. Biological monitoring of exposure to chemicals: metals. New York: John Wiley and Sons; 1991. p. 175–96.
- [9] Dardenne M. Zinc and immune function. Eur J Clin Nutr 2002;56:S20-3.
- [10] Sahaphong S, Linthong V, Wanichanon C, Riengrojpitak S, Kangwanrangsan N, Viyanant V, et al. Morphofunctional study of the hemocytes of *Haliotis asinina*. J Shellfish Res 2001;20:711–6.
- [11] Galloway TS, Depledge MH. Immunotoxicity in invertebrates: measurement and ecotoxicological relevance. Ecotoxicology 2001;10:5–23.
- [12] Hooper C, Day R, Slocombe R, Handlinger J, Benkendorff K. Stress and immune responses in abalone: limitations in current knowledge and investigative methods based on other models. Fish Shellfish Immunol 2007;22:363–79.
- [13] Travers MA, Mirella Da Silva P, Le Goïc N, Marie D, Donval A, Huchette S, et al. Morphologic, cytometric and functional characterisation of abalone (*Haliotis tuberculata*) haemocytes. Fish Shellfish Immunol 2008:24:400–11.
- [14] Fagotti A, Di Rosa I, Simoncelli F, Pipe RK, Panara F, Pascolini R. The effects of copper on actin and fibronectin organization in *Mytilus galloprovinciallis* haemocytes. Dev Comp Immunol 1996;20:383–91.
- [15] Olabbarietta I, L'Azou B, Yuric S, Cambar J, Cajaraville MP. In vitro effects of cadmium on two different animal cell models. Toxicol In Vitro 2001;15:511–7.
- [16] Sauvé S, Brousseau P, Pellerin J, Morin Y, Senécal L, Goudreau P, et al. Phagocytic activity of marine and freshwater bivalves: in vitro exposure of hemocytes to metals (Ag, Cd, Hg and Zn). Aquat Toxicol 2002;58:189–200.
- [17] Gagnaire B, Thomas-Guyon H, Renault T. In vitro effects of cadmium and mercury on pacific oyster, *Crassostrea gigas* (Thunberg), haemocytes. Fish Shellfish Immunol 2004;16:501–12.
- [18] Duchemin MB, Auffret M, Wessel N, Fortier M, Morin Y, Pellerin J, et al. Multiple experimental approaches of immunotoxic effects of mercury chloride in the blue mussel, *Mytilus edulis*, through in vivo, in tubo and in vitro exposures. Environ Pollut 2008;153:416–23.
- [19] Cheng W, Hsiao IS, Hsu CH, Chen JC. Change in water temperature on the immune response of Taiwan abalone *Haliotis diversicolor supersexta* and its susceptibility to *Vibrio parahaemolyticus*. Fish Shellfish Immunol 2004;17:235–43.
- [20] Cheng W, Juang FM, Chen JC. The immune response of Taiwan abalone Haliotis diversicolor supersexta and its susceptibility to Vibrio parahaemolyticus at different salinity levels. Fish Shellfish Immunol 2004;16:295–306.
- [21] Travers MA, Le Goïc N, Huchette S, Koken M, Paillard C. Summer immune depression associated with increased susceptibility of the European abalone *Haliotis tuberculata* to *Vibrio harveyi* infection. Fish Shellfish Immunol 2008;25:800–8.
- [22] Wang SH, Wang YL, Zhang ZX, Ralph J, Weng ZH, Zou ZH, et al. Response of innate immune factors in abalone *Haliotis diversicolor supersexta* to pathogenic or non-pathogenic infection. J Shellfish Res 2004;23:1173–7.
- [23] Martello LB, Friedman CS, Tjeerdema RS. Combined effects of pentachlorophenol and salinity stress on phagocytic and chemotactic function in two species of abalone. Aquat Toxicol 2000;49:213–25.
- [24] Lebel J-M, Giard W, Favrel P, Boucaud-Camou E. Effects of different vertebrate growth factors on primary cultures of hemocytes from the gastropod mollusc, *Haliotis tuberculata*. Biol Cell 1996;86:67–72.
- [25] Poncet J-M, Serpentini A, Thiébot B, Villers C, Bocquet J, Boucaud-Camou E, et al. In vitro synthesis of proteoglycans and collagen in primary cultures of mantle cells from the nacreous mollusk, *Haliotis tuberculata*: a new model for study of molluscan extracellular matrix. Mar Biotechnol 2000;2:387–98.
- [26] Serpentini A, Ghayor C, Hebert V, Galéra P, Pujol J-P, Boucaud-Camou E, et al. De novo synthesis and identification of collagen transcripts in hemocytes from the gastropod mollusc, Haliotis tuberculata. J Exp Zool 2000;287:275–84.
- [27] Riinkevich B. Marine invertebrate cell cultures: new millennium trends. Mar Biotechnol 2005;7:429–39.
- [28] Le Pennec G, Le Pennec M. Acinar primary cell culture from the digestive gland of *Pecten maximus* (L.): an original model for ecotoxcological purposes. Exp Mar Biol Ecol 2001;259:171–87.
- [29] Schirmer K. Proposal to improve vertebrate cell cultures to establish them as substitutes for the regulatory testing of chemicals and effluents using fish. Toxicology 2006;224:163–83.
- [30] Wu JP, Chen HC, Huang DJ. Histopathological alterations in gills of white shrimp, *Litopenaeus vannamei* (Boone) after acute exposure to cadmium and zinc. Bull Environ Contam Toxicol 2009;82:90–5.
- [31] Stabili L, Pagliara P. Effect of zinc on lusozyme-like activity of the seastar Marthasterias glacialis (Echinodermata Asteroidea) mucus. J Invertebr Pathol 2009;100:189–92.
- [32] Farcy E, Serpentini A, Fiévet B, Lebel J-M. Identification of cDNAs encoding HSP70 and HSP90 in the abalone *Haliotis tuberculata*: transcriptional induction in response to thermal stress in hemocyte primary culture. Comp Biochem Physiol B Biochem Mol Biol. 2007;146:540–50.
- [33] Bachère E, Chagot D, Grizel H. Separation of Crassostrea gigas hemocytes by density gradient centrifugation and counterflow centrifugal elutriation. Dev Comp Immunol 1988;12:549–59.

- [34] Gorski J, Nugegoda D. Toxicity of trace metals to juvenile abalone, Haliotis rubra following short-term exposure. Bull Environ Contam Toxicol 2006;77:732–40.
- [35] Mosmann T. Rapid colorimetric assay for cellular growth and survival: application to proliferation and cytotoxicity assays. J Immunol Methods 1983;65:55–63.
- [36] Poncet J-M, Serpentini A, Boucaud-Camou E, Lebel J-M. Cryopreservation of mantle dissociated cells from *Haliotis tuberculata* (Gastropoda) and postthawed primary cells cultures. Cryobiology 2002;44:38–45.
- [37] Bass DA, Parce JW, Dechatelet LR, Szejda P, Seeds MC, Thomas M. Flow cytometric studies of oxidativeproduct formation by neutrophils: a graded response to membrane stimulation. J Immunol 1983;130:1910–7.
- [38] Lambert C, Soudant P, Choquet G, Paillard C. Measurement of *Crassostrea gigas* hemocyte oxidative metabolism by flow cytometry and the inhibiting capacity of pathogenic vibrios. Fish Shellfish Immunol 2003;15:225–40.
- [39] Delaporte M, Soudant P, Moal J, Lambert C, Quere C, Miner P, et al. Effect of a mono-specific algal diet on immune functions in two bivalves species Crassostrea gigas and Ruditapes Philippinarum. J Exp Biol 2003;206:3053-64.
- [40] Hégaret H, Wikfors GH, Soudant P. Flow cytometric analysis of haemocytes from eastern oysters, *Crassostrea virginica*, subjected to a sudden temperature elevation: II. Haemocyte functions: aggregation, viability, phagocytosis, and respiratory burst. J Exp Mar Biol Ecol 2003;293:249–65.
- [41] Labreuche Y, Soudant P, Gonçalves M, Lambert C, Nicolas JL. Effect of extracellular products from the pathogenic Vibrio aestuarianus strain 01/32 on lethality and cellular immune responses of the oyster Crassostrea gigas. Dev Comp Immunol 2006;30:367–79.
- [42] González F, Fárez-Vidal ME, Arias JM, Montoya E. Partial purification and biochemical properties of acid and alkaline phosphatases from *Myxococcus coralloides* D. J Appl Bacteriol 1994;77:567–73.
- [43] Hernandez-Lopez J, Gollas-Galvan T, Gomez-Jimenez S, Portillo-Clark G, Vargas-Albores F. In the spiny lobster (*Panulirus interruptus*) the prophenoloxidase is located in plasma not in haemocytes. Fish Shellfish Immunol 2003;14:105–14.
- [44] Bradford MM. A refined and sensitive method for quantification of microgram quantities of proteins utilizing the principl of protein-dye binding. Anal Biochem 1976;72:248–54.
- [45] Brousseau P, Pellerin J, Morin Y, Cyr D, Blakley B, Boermans H, et al. Flow cytometry as a tool to monitor the disturbance of phagocytosis in the clam *Mya arenaria* hemocytes following *in vitro* exposure to heavy metals. Toxicology 2000;142:145–56.
- [46] Song J, Lee SC, Kim SS, Koh HJ, Kwon OW, Kang JJ, et al. Zn²⁺-induced cell death is mediated by the induction of intracellular ROS in ARPE-19 cells. Curr Eye Res 2004;28:195–201.
- [47] Gómez-Mendikute A, Cajaraville MP. Comparative effects of cadmium, copper, paraquat and benzo[a]pyrene on the actin cytoskeleton and production of reactive oxygen species (ROS) in mussel haemocytes. Toxicol In Vitro 2003;17:539–46.
- [48] Larson KG, Roberson BS, Hetrick FM. Effect of environmental pollutants on the chemiluminescence of hemocytes from the American oyster *Crassostrea virginica*. Dis Aquat Org 1989;6:131–6.
- [49] Beckmann N, Morse MP, Moore CM. Comparative study of phagocytosis in normal and diseased hémocytes of the bivalve mollusc *Mya arenaria*. J Invertebr Pathol 1992;59:124–32.
- [50] Cima F, Marin MG, Matozzo V, Da Ros L, Ballarin L. Biomarkers for TBT immunotoxycity studies on the cultivated clam *Tapes philippinarum* (Adams and Reeve, 1850). Mar Pollut Bull 1999;39:112–5.
- [51] Kaloyianni M, Ragia V, Tzeranaki I, Dailianis S. The influence of Zn on signaling pathways and attachment of *Mytilus galloprovincialis* haemocytes to extracellular matrix proteins. Comp Biochem Physiol C 2006;144:93–100.
- [52] Thiagarajan R, Gopalakrishnan S, Thilagam H. Immunomodulation in the marine green mussel *Perna viridis* exposed to sub-lethal concentrations of Cu and Hg, Arch Environ Contam Toxicol 2006;51:393–9.
- [53] Bado-Nilles A, Gagnaire B, Thomas-Guyon H, Le Floch S, Renault T. Effects of 16 pure hydrocarbons and two oils on haemocyte and haemolymphatic parameters in the Pacific oyster, *Crassostrea gigas*. Toxicol In Vitro 2008;22:1610–7.
- [54] Gagnaire B, Thomas-Guyon H, Burgeot Th, Renault T. Pollutant effects on Pacific oyster, *Crassostrea gigas* (Thunberg), hemocytes: screening of 23 molecules using flow cytometry. Cell Biol Toxicol 2006;22:1–14.
- [55] Mazón LI, Gonzalez G, Vicario A, Estomba A, Aguirre A. Inhibition of esterase in the marine gastropod *Littorina littorea* exposed to cadmium. Ecotoxicol Environ Saf 1998;41:284–7.
- [56] Zvereva E, Serebrov V, Glupov V, Dubovskiy I. Activity and heavy metal resistance of non-specific esterase in leaf beetle *Chrysomela lapponica* from polluted and unpolluted habitats. Comp Biochem Physiol C 2003;135:383–91.
- [57] Vallee BL, Falchuk KH. The biochemical basis of zinc physiology. Phys Rev 1993;73:79–118.
- [58] Coles JA, Pipe RK. Phenoloxidase activity in the haemolymph and haemocytes of the marine mussel *Mytilus edulis*. Fish Shellfish Immunol 1994;4:337–52.
- [59] Liu G, Yang L, Fan T, Cong R, Tang Z, Sun W, et al. Purification and characterization of phenoloxidase from crab *Charybdis japonica*. Fish Shellfish Immunol 2006;20:47–57.
- [60] Lacoue-Labarthe T, Bustamante P, Hörlin E, Luna-Acosta A, Bado-Nilles A, Thomas-Guyon H. Phenoloxidase activation in the embryo of the common cuttlefish Sepia officinalis and responses to the Ag and Cu exposure. Fish Shellfish Immunol 2009;27:516–21.

- [61] Tujula N, Radford J, Nair SV, Raftos DA. Effects of tributyltin and other metals on the phenoloxidase activating system of the tunicate, *Styela plicata*. Aquat Toxicol 2001;55:191–201.
- [62] Mazorra MT, Rubio JA, Blasco J. Acid and alkaline phosphatase activities in the clam *Scrobicularia plana*: kinetic characteristics and effects of heavy metals. Comp Biochem Physiol B 2002;131:241–9.
- [63] Jing G, Li Y, Xie L, Zhang R. Metal accumulation and enzyme activities in gills and digestive gland of pearl oyster (*Pinctada fucata*) exposed to copper. Comp Biochem Physiol C 2006;144:184–90.
- [64] Rajalakshmi S, Mohandas A. Copper-induced changes in tissue enzyme activity in a freshwater mussel. Ecotoxicol Environ Saf 2005;62:140–3.
- [65] Shuhong W, Yilei W, Zhaoxia Z, Jack R, Zhaohong W, Zhihua Z, et al. Response of innate immune factors in abalone *Haliotis diversicolor supertexta* to pathogenic or non-pathogenic infection. J Shellfish Res 2004;23:1173–7.
- [66] Nakayama K, Maruyama T. Differential production of active oxygen species in photo-symbiotic and non-symbiotic bivalves. Dev Comp Immunol 1998;22:151–9.